

# Hydroxyiminodisuccinic acid (HIDS): A novel biodegradable chelating ligand for the increase of iron bioavailability and arsenic phytoextraction

著者	Rahman M. Azizur, Hasegawa Hiroshi, Kadohashi K., Maki Teruya, Ueda Kazumasa
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1       **Hydroxyiminodisuccinic acid (HIDS): A Novel Biodegradable**  
2       **Chelating Ligand for the Increase of Iron Bioavailability and**  
3               **Arsenic Phytoextraction**

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6       **M. Azizur Rahman\***; **H. Hasegawa\***; **K. Kadohashi**; **T. Maki**; **K. Ueda**

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10  
11       Graduate School of Natural Science & Technology, Kanazawa University, Kakuma, Kanazawa  
12                               920-1192, Japan

13  
14  
15  
16  
17  
18                               \*Corresponding authors

19                               E-mail: hhiroshi@t.kanazawa-u.ac.jp (H. Hasegawa)

20                               aziz\_ju@yahoo.com (M. Azizur Rahman)

21                               Tel/Fax: 81-76-234-4792

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23

**Abstract**

The influence of biodegradable chelating ligands on arsenic and iron uptake by hydroponically grown rice seedlings (*Oryza sativa* L.) was investigated. Even though the growth solution contained sufficient Fe, the growth of rice seedlings gradually decreased up to 76% with the increase of pH of the solution from 7 to 11. Iron forms insoluble ferric hydroxide complexes at neutral or alkaline pH in oxic condition. Chelating ligands produce soluble 'Fe-ligand complex' which assist Fe uptake in plants. The biodegradable chelating ligand hydroxyiminodisuccinic acid (HIDS) was more efficient than those of ethylenediaminetetraacetic acid (EDTA), ethylenediaminedisuccinic acid (EDDS), and iminodisuccinic acid (IDS) in the increase of Fe uptake and growth of rice seedling. A total of  $79\pm 20$ ,  $87\pm 6$ ,  $116\pm 15$ , and  $63\pm 18$  mg dry biomass of rice seedlings were produced with the addition of 0.5 mM of EDDS, EDTA, HIDS, and IDS in the nutrient solution, respectively. The Fe concentrations in rice tissues were  $117\pm 15$ ,  $82\pm 8$ ,  $167\pm 25$ , and  $118\pm 22$   $\mu\text{mol g}^{-1}$  dry weights when 0.25 mM of EDDS, EDTA, HIDS, and IDS were added to the nutrient solution, respectively. Most of the Fe accumulated in rice tissues was stored in roots after the addition of chelating ligands in the solution. The results indicate that the HIDS would be a potential alternative to environmentally persistent EDTA for the increase of Fe uptake and plant growth. The HIDS also increased As uptake in rice root though its translocation from root to shoot was not augmented. This study reports HIDS for the first time as a promising chelating ligand for the enhancement of Fe bioavailability and As phytoextraction.

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**Keywords:** Arsenic, Iron, Chelating ligands, Rice (*Oryza sativa* L.), Hydroponics, Bioavailable,

HIDS.

47

## 48 **1. Introduction**

49 Iron is an essential micronutrient for plants, which plays important roles in respiration,  
50 photosynthesis, and many other cellular functions such as DNA synthesis, nitrogen fixation, and  
51 hormone production (Vert et al., 2002). Although abundant in nature it forms insoluble ferric  
52 hydroxide complexes (also known as Fe-plaque) at neutral or alkaline pH in oxic condition  
53 (Guerinot and Yi, 1994). The Fe-plaque formation in the rhizosphere soils, however, results in  
54 the Fe deficiency to plants. In nature, rhizospheric microbes exude siderophores to the root-  
55 plaque interface. These siderophores solubilize ferric iron in the rhizosphere, render its  
56 bioavailability, and plants uptake the Fe by specific membrane receptors (Romheld, 1987).

57 Elevated levels of As in soil from natural and anthropogenic sources is a threat to plants'  
58 health (Rahman et al., 2008). Remediation of contaminated soil is important to prevent As  
59 deposition in food crops and its subsequent transfer into the human body through the food chains  
60 (Rahman et al., 2008). Phytoremediation becomes a promising alternative and environmentally  
61 safe technology for the remediation of environmental pollutants (Raskin et al., 1997; Tu et al.,  
62 2002). An essential prerequisite for phytoremediation of contaminated soil is solubility and  
63 bioavailability of As (Fitz and Wenzel, 2002). But the solubility and bioavailability of As  
64 becomes reduced by adsorption to variable charged minerals (Fe and Al) at alkaline pH (Xu et al.,  
65 2008). In the past decade, chelant-enhanced phytoremediation has received much attention  
66 (Pastor et al., 2007). This technique aims to cleanse polluted soils by solubilizing the toxic  
67 metals, allowing it to be accumulated in plants that would subsequently remove toxic metal from  
68 the site. Publications on chelant-enhanced phytoremediation have increased steadily to about 15-  
69 20 per year in the last few years, indicating that this is a growing and active research field  
70 (Nowack et al., 2006).

71 Research on the interaction of plants with chelating ligands started in the 1950s with a  
72 view to reduce the deficiencies of the essential nutrients such as Fe, Mn, Cu, and Zn (Wenger et  
73 al., 2005). Among all soil-applied Fe fertilizers, synthetic Fe(III)-chelates, mainly Fe(III)-  
74 chelates of polyaminecarboxylic acids with phenolic groups, such as ethylenediamine di(*o*-  
75 hydroxyphenylacetic) acid (EDDHA), and ethylenediamine di(2-hydroxy-4-methylphenylacetic)  
76 acid, are the most effective and commonly used (Alvarez-Fernandez et al., 2005). On the other  
77 hand reports on As phytoextraction by chelating ligands is limited though a number of  
78 investigations have been conducted on chelant-enhanced phytoextraction of Pb, Zn, Hg, Cu and  
79 some other heavy metals (Luo et al., 2005). Ethylenediaminetetraacetic acid (EDTA) has been  
80 very popular to achieve this purpose, but it is quite persistent in the environment because of its  
81 low biodegradability. This, in combination with its high affinity for heavy metal complexation,  
82 results in an increased risk of leaching. EDTA also impairs plant growth severely, even at low  
83 concentrations (Bucheli-Witschel and Egli, 2001).

84 Biodegradable chelating ligands, such as ethylenediaminedisuccinic acid (EDDS),  
85 Hydroxyiminodisuccinic acid (HIDS), and iminodisuccinic acid (IDS) would be good choice and  
86 alternative to less biodegradable EDTA. The physicochemical properties of EDDS, EDTA, and  
87 IDS have already been discussed and tested for the phytoextraction of heavy metals by a number  
88 of researchers (Helena et al., 2003; Evangelou et al., 2007). HIDS is a new chelating ligand  
89 introduced by Nippon Shokubai Co. Ltd. It is one of the highly biodegradable (biodegradation  
90 rate is about 22.4% within 48 h) and safe chelating ligands. It traps and inactivates various kinds  
91 of metals ions over a wide range of pH, particularly  $\text{Fe}^{3+}$  and  $\text{Cu}^{2+}$ , as well as  $\text{Ca}^{2+}$  and  $\text{Mg}^{2+}$ ;  
92 shows high stability in harsh conditions and high temperature (80 °C); is highly soluble in  
93 aqueous alkaline solution (Sokubai, 2009). Because of high degradation rate and high stability  
94 constant with  $\text{Fe}^{3+}$  ( $\text{pK}_a\text{Fe}^{3+}$  is 12.5) of HIDS, we become interested to investigate the  
95 effectiveness of the chelating ligand for the increase of Fe bioavailability and phytoremediation

96 of As. The EDTA, EDDS, and IDS were also used in the present study to compare the results of  
97 HIDS. Our research approach was to find a biodegradable and eco-friendly chelating ligand that  
98 is more desirable than EDTA or EDDS for Fe bioavailability and As phytoextraction.

99

## 100 **2. Materials and Methods**

### 101 *2.1. Seed sterilization*

102 Rice seeds of BRRI dhan 29 were collected from Bangladesh Rice Research Institute.  
103 The seeds were surface-sterilized before using them in the experiment. For sterilization, about  
104 100 g seeds were soaked in 200 mL of 1% methyl-1-butylcarbamoyl-2-benzimidazole carbonate  
105 solution for 10 min. After that, the seeds were washed by deionized (DI) water (using an E-pure  
106 system (Barnstead)) and kept in DI water at 20 °C for 24 h. The seeds were then washed and  
107 transferred to DI water of 45 °C for 2 min, and of 52 °C for 10 min.

108

### 109 *2.2. Chemicals*

110 Stock solutions of EDTA, EDDS, HIDS, and IDS were prepared by dissolving  
111 ethylenediamine-N,N,N',N'-tetraacetic acid (Dojindo Molecular Technologies, Japan),  
112 ethylenediamine-N, N'-disuccinic acid (Chelest), tetrasodium 3-hydroxy-2,2'-iminodisuccinate  
113 (Nippon Syokubai, Japan), and tetrasodium iminodisuccinate (Bayer) in 0.1 M sodium hydroxide,  
114 respectively. Other reagents were of analytical grade or better. All solutions were prepared with  
115 DI water.

116

### 117 *2.3. Nutrient solution*

118 Sterilized rice seeds were germinated on pre-sterilized bloating paper (seed bed) with  
119 standard murashige and skoog (MS)([Murashige and Skoog, 1962](#)). Iron concentration in the

120 experimental solution was 0.36 mM while its concentration was 27.8 mg L<sup>-1</sup> in pre-experimental  
121 solution (used for growing rice seedling prior to the experiment). The pH of the pre-experimental  
122 solution was adjusted to 6.5 while the pH of experimental solution was 9.0. Rice seedlings were  
123 grown on the seed bed for 1 wk. In preparing MS culture solution, FeSO<sub>4</sub>·7H<sub>2</sub>O was used as Fe  
124 source instead of NaFe(III)-EDTA.

125

#### 126 *2.4. Experimental setup*

127 Rice seedlings were transferred to the experimental solution after one week of growth in  
128 pre-experimental solution. In the experimental solution, rice seedlings were grown in two steps.  
129 In the first step, rice seedlings were grown with different concentrations of chelating ligands (up  
130 to 2.50 mM) to observe the effect of chelating ligands on Fe uptake. In the second step, 6.0 μM  
131 of As (Na<sub>2</sub>HAsO<sub>4</sub>·7H<sub>2</sub>O) was added to the nutrient solutions containing 1.0 mM of chelating  
132 ligands to see the effect of chelating ligands on Fe and As uptake. Iron concentration in the  
133 experimental solution was 0.36 mM, and the pH of the solution was adjusted to 9 using 0.1 M  
134 KOH. About 100 mL of the solution was taken into 250-mL polystyrene bottles with three  
135 replications, and three uniform seedlings were cultivated in each bottle. The experiment was  
136 performed following randomized design. Rice plants were grown in a plant growth chamber and  
137 the conditions in the chamber were set as 14:10 h light/dark schedule, 100-125 μ E m<sup>-2</sup> s<sup>-1</sup> light  
138 intensity, 22(±2) °C temperatures. Rice seedlings were grown in experimental solution for 5 d.

139

#### 140 *2.5. CBE-extraction of Fe-plaques*

141 At harvest, the shoots were cut from 1 cm above the roots and separated. The Fe-plaques  
142 from root surfaces were extracted using citrate-bicarbonate-ethylenediaminetetraacetate (CBE)-  
143 technique, a modified method of dithionite-citrate-bicarbonate extraction by [Taylor and Crowder](#)  
144 [\(1983\)](#) to determine the real amount of Fe and As contents in rice tissues. The CBE solution was

145 prepared from 0.03, 0.125 and 0.050 M of sodium citrate, sodium bicarbonate, and EDTA,  
146 respectively. Roots were treated with 30 mL of CBE solution for 60 min at room temperature.  
147 The roots were then rinsed with deionized water for 3 times, and the rinsed water was added to  
148 the CBE-extracts to make a total of 30 mL.

149

## 150 *2.6. Sample preparation*

151 After rinsing with deionized water for four times, the root samples were kept on clean  
152 absorbent paper to remove the water from the root surfaces. Both the root and the shoot samples  
153 were dried at 65 °C until they reached in a constant weight. Then the dried samples were  
154 weighted and taken into 50-mL polyethylene tubes for digestion. Five mL of 65% HNO<sub>3</sub> were  
155 added to the sample and kept for 12 h. The samples were heated on a heating block at 95 °C for 2  
156 h. After cooling to room temperature, 3 mL of 30% hydrogen peroxide were added, and the  
157 samples were heated again at 105 °C for 20 min. Then, the digests were diluted to 30 mL with DI  
158 and analyzed for As and Fe.

159

## 160 *2.7. Chemical analysis*

161 Arsenic and Fe were analyzed using graphite-furnace atomic absorption spectrometer (Z-  
162 8100, Hitachi, Japan). Certified standard reference material 1573a (tomato leaf from NIST,  
163 USA) was used to check the accuracy of analysis. Arsenic concentration in certified standard  
164 reference materials was  $0.112 \pm 0.004 \mu\text{g g}^{-1}$  dry weight (all the reported data in this article are  
165 expressed as dry weight) while the measured concentration was  $0.114 \pm 0.002 \mu\text{g g}^{-1}$ . The  
166 concentrations detected in all samples were above the instrumental limits of detection ( $\geq 0.01$   
167  $\mu\text{M}$  in water sample).

168 All chemical reagents used in this experiment were of analytical grade. Glassware and  
169 dishes were washed with detergent and 1 N HCL solution, and rinsed with DI water for eight



170 times before use. In each analytical batch, at least two reagent blanks and three replicate samples  
171 were included.

172

### 173 **3. Results and Discussions**

#### 174 *3.1. Effect of pH on rice growth*

175 Rice seedlings were grown in nutrient solution adjusted to different pH ranging between  
176 6 and 11. Results show that the biomass production of rice seedlings was affected by the pH  
177 significantly. The highest biomass of rice seedling ( $83\pm 7$  mg) was observed at pH 7, which was  
178 about 16, 19, 43, and 76% higher than those at pH 8, 9, 10, and 11, respectively (Fig. 1). The rice  
179 growth remain unchanged, and even died at pH 10 and 11. Rice plants have a tendency of higher  
180 Fe uptake than that of other plants (Becker and Asch, 2005). But the pH of the growth medium  
181 plays an important role in Fe bioavailability and uptake. Even though the Fe is sufficient in  
182 growth medium, it forms insoluble ferric hydroxide complexes at alkaline pH in oxic condition  
183 (Cohen et al., 1998). Therefore, Fe bioavailability and uptake decreases drastically. In the present  
184 study, it was observed that the Fe concentrations in tissues of rice seedlings were highest at pH 7  
185 compared to those at other pHs (Fig. 2). This trend of Fe uptake in rice tissues is correlated to  
186 that of biomass production of rice seedlings (Fig. 1). The result implies that the influence of pH  
187 on rice growth is the ultimate effect of reduced Fe bioavailability and uptake. Moreover, Fe  
188 concentrations on root surfaces of rice seedlings were lowest at pH 7 and 8 compared to those at  
189 other pHs (Fig. 2). High level of Fe on root surfaces of rice seedling at pH 11 reveals the  
190 formation of Fe-hydroxides (Fe-plaque) on root surfaces, which decreased the Fe uptake in rice  
191 tissues. Formation of Fe-plaques on the roots of wetland plants (Hansel et al., 2001) and  
192 hydroponically grown rice seedling (Hu et al., 2005) have also been reported. The precipitation  
193 of ferric (oxyhydro)-oxides ( $\text{FeO}_x$ ) and its association with phytoplankton surfaces, both in

194 natural conditions and laboratory cultures, has been reported by [Tang and Morel \(2006\)](#).  
195 [Robinson et al. \(2006\)](#) also found the occurrence of Fe-plaque on aquatic macrophytes collected  
196 from the Taupo Volcanic zone, New Zealand.

197 The Fe deficiency results in Fe-chlorosis in green leaves, which retards plant growth, and  
198 leads to the reduction of crop yields ([Guerinot and Yi, 1994](#)). The results of the present study  
199 also reveal that the growth of rice seedling decreased drastically at higher pH, which is the  
200 consequence of Fe-chlorosis.

201

### 202 *3.2. Influence of chelating ligands on Fe uptake-translocation*

203 Influence of EDDS, EDTA, HIDS, and IDS on Fe uptake and translocation in rice  
204 seedlings were investigated at different concentrations of the ligands ranging between 0.1 and  
205 2.5 mM. Results showed that Fe uptake in rice seedling differed significantly with the type and  
206 concentrations of the chelating ligands. Iron uptake was highest at 0.25 mM of the chelating  
207 ligands compared to the control treatment. Iron uptake decreased gradually with the increase of  
208 chelating ligand concentrations above 0.25 mM ([Fig. 3](#)). The effectiveness of HIDS and EDDS  
209 in the increase of Fe uptake in rice tissues was higher than that of EDTA and IDS. Iron  
210 concentrations in roots of rice seedling were  $35\pm 3$  and  $44\pm 2$   $\mu\text{mol g}^{-1}$  when the HIDS  
211 concentrations in the nutrient solution were 0.10 and 0.25 mM, respectively. These  
212 concentrations were significantly higher than those of other chelating ligands.

213 Iron concentrations in shoots of rice seedlings were significantly lower than those in roots,  
214 and were about identical up to 0.25 mM of chelating ligand treatment. Iron content in shoots  
215 decreased with the gradual increase of chelating ligands from 0.25 to 2.50 mM ([Fig. 3](#)). The  
216 results indicate that the translocation of Fe from roots to shoots was not affected by lower dose  
217 of the chelating ligands. The translocation of Fe was inhibited by the chelating ligands at higher  
218 doses ( $> 0.25$  mM).

219 Although abundant in nature, Fe is often unavailable to plants, especially at neutral or  
220 alkaline pH, because of the formation of insoluble ferric hydroxide complexes in oxic condition  
221 (Robinson et al., 2006). Precipitation of Fe in the rhizosphere, however, may result in the Fe  
222 deficiency to the plants. Chelating ligands are used in agriculture as additives in micronutrient  
223 fertilizers for the increase of Fe bioavailability. Although some chelating ligands have been  
224 reported to increase Fe uptake/translocation in plant, inhibition of Fe uptake/translocation by  
225 ligands has also been reported. Chaney et al. (1972) reported that  
226 bathophenanthrolinedisulfonate (BPDS) was the most effective inhibitor of Fe  
227 uptake/translocation, followed by EDTA > DTPA (diethylenetriaminepentaacetic acid) > CDTA  
228 (diaminocyclohexanetetraacetic acid) >> EDDHA. The BPDS inhibited  $^{59}\text{Fe}$  movement to the  
229 exudate by 99.7% even at the lowest level of competitor. The BPDS inhibits Fe translocation by  
230 10-100 times compared to those of EDTA, DTPA, or CDTA. Chaney et al. (1972) also observed  
231 that EDDHA, the chelator with the highest  $\text{Fe}^{3+}$  stability constant, only slightly inhibited or  
232 actually promoted Fe uptake/translocation, whereas the BPDS with the highest  $\text{Fe}^{2+}$  stability  
233 constant was a severe inhibitor. Thus, stability constant of Fe-ligand ( $\log K_{\text{FeL}}$ ) would be one of  
234 the important determinants for the promotion or inhibition of Fe uptake/translocation.

### 235 3.3. Effect of chelating ligands on rice growth

236 Rice seedlings were grown in alkaline nutrient solution (pH 9) containing 0.10, 0.25, 0.50,  
237 1.00, and 2.50 mM of chelating ligands and 0.36 mM of Fe. Results show that the growth of rice  
238 seedlings was increased with the increase of HIDS and EDTA concentrations up to 1.0 mM, and  
239 the growth was decreased at 2.5 mM of chelating ligand concentrations (Fig. 4). The highest  
240 biomass production ( $141 \pm 21$  mg) of rice was observed when 1.0 mM of EDTA was added to the  
241 nutrient solution followed by  $127 \pm 8$ ,  $82 \pm 19$ , and  $75 \pm 4$  mg for HIDS, EDDS, and IDS,  
242 respectively.

243 Chelating ligands have been used to enhance Fe bioavailability (Alvarez-Fernandez et al.,  
244 2005). The concentration of chelating ligands in the nutrient medium is important for the  
245 solubilization of precipitated Fe and the increase of its bioavailability. In the present study, it was  
246 observed that the rice seedling produce highest biomass at 1.0 mM chelating ligand  
247 concentrations, and the growth remain unchanged, and even died at higher concentration (>1.0  
248 mM).

249 Although the growth of all organisms is dependent on the acquisition of the proper  
250 quantities of trace elements, excess amount of some metals such as Fe, zinc, manganese, and  
251 copper produce toxic effects (Morel and Hering, 1993). However, ferric ions and their complexes,  
252 which have low solubility in aquatic system, are extensively buffered by chelation (Morel and  
253 Hering, 1993), and increase their dissolved concentration. The dissolved concentration of Fe  
254 determines its rate of uptake by the organisms. Anderson and Morel (1982) reported that the Fe  
255 uptake rate in laboratory cultures of the marine diatom *Thalassosira weissflogii* is a unique  
256 function of the free ferric ion concentration at the presence of  $10^{-5}$  M of various chelating ligands  
257 ( $1.4 \times 10^7$  cells  $L^{-1}$ ). Hudson and Morel (1990) reported that in Fe-limited culture of marine  
258 diatom *Thalassosira weissflogii* ( $10^7$  cells  $L^{-1}$ ) containing  $10^{-8}$  M Fe and  $10^{-5}$  M EDTA and with  
259 white-light illumination, both the thermal dissociation of FeEDTA and its photoreduction and  
260 reoxidation contribute to the formation of the dissolved inorganic Fe(III) pool responsible for the  
261 Fe uptake. In this case, growth of rice seedlings was inhibited by the free ferric ion that was  
262 increased by the addition of higher level of chelating ligands.

263 Toxicity of chelating ligands on plants has not been studied extensively. So, it is difficult  
264 to interpret the direct toxicity of chelating ligands on plants. Since most of the chelating ligands  
265 are synthetic compounds, no nutrient carriers in the plasma membrane are thought to exist  
266 (Berne and Levy, 1998). Also, synthetic chelates cannot slip through the plasma membrane as  
267 they are too large and polar to move through the plasma lemma lipid bilayer (Berne and Levy,

268 1998). Tanton and Crowdy (1972) observed that most solutes moved into some endodermal  
269 passage cells adjacent to the casparian strip intracellularly to the other side of the strip, and then  
270 extracellularly to the xylem. The passage cells may include the aquaporins and there may be  
271 selectivity toward molecules. Paul et al. (2003) reported that Swiss chard uptakes a considerable  
272 amount of EDTA from chelator-buffered hydroponic solution through transpirational flow that  
273 occurs via apoplastically.

274

### 275 *3.4. Influence of chelating ligands and As on rice growth*

276 Chelating ligand treated rice seedlings were grown with and without As to investigate the  
277 effect of As and chelating ligands on rice growth. Results show that As does not have a  
278 consistent effect on rice growth as chelating ligand has. Rice growth was not affected by As  
279 when chelating ligand was not treated. The highest growth of rice seedling was observed in  
280 HIDS treated medium. The inconsistent effect of chelating ligand and As on rice growth suggest  
281 that in the presence of chelating ligands lower level of As in the growth medium does not affect  
282 rice growth significantly. It has been reported that rice growth is not affected by low level of As  
283 though the growth decrease drastically with the increase of As in the soil. Abedin and Meharg  
284 (2002) also reported that low level of As in water (about 2.0 mg L<sup>-1</sup>) does not show toxicity to  
285 both rice germination and rice growth, but the rice germination and growth were adversely  
286 affected by higher As level.

287

### 288 *3.5. Influence of chelating ligands and As on Fe uptake/translocation*

289 Iron uptake in rice seedling was affected by chelating ligands and As significantly. Iron  
290 concentration was measured both in root surfaces and plant tissues. Results show that the Fe  
291 concentration was higher in rice root surfaces of control treatment (without chelating ligands)  
292 while its concentration was higher in plant tissues of ligand treated nutrient solution (Fig. 5). The

293 highest Fe contents were found in tissues of rice seedlings treated with EDTA or HIDS and As.  
294 Increasing Fe uptake by chelating ligands, especially EDTA and HIDS, can be explained by the  
295 adsorption of As(III)-EDTA/-HIDS complex on the Fe-plaques of rice root surfaces and  
296 dissociation of the complex to release of Fe(III)-EDTA/-HIDS into solution. The release of  
297 Fe(III)-EDTA/-HIDS into the culture solution results in the increase of Fe uptake. Adsorption of  
298 metal-EDTA to the surface of Fe oxides and dissociation of the complex and release of Fe(III)-  
299 EDTA has been reported by [Nowack and Sigg \(1997\)](#).

300 Strong ligands, such as EDTA, complex with metals in natural systems. Adsorption of  
301 uncomplexed EDTA on metal oxides (Fe-oxides, Al-oxides) has been studied previously  
302 ([Bowers and Huang, 1985](#); [Blesa et al., 2000](#)). The EDTA has been reported to exist as complex  
303 species of metals (mainly CaEDTA, ZnEDTA, and Fe(III)EDTA) in natural waters ([Xue et al.,](#)  
304 [1995](#)). Dissolution reactions of Fe-oxides in the presence of metal-EDTA complexes have also  
305 been reported by [Nowack and Sigg \(1997\)](#).

306

### 307 *3.6. Arsenic uptake/translocation affected by chelating ligands*

308 Arsenic contents in roots, shoots, and root surfaces of rice seedling were determined to  
309 assess the effect of chelating ligands on As uptake. Results show that As was stored mostly in  
310 roots followed by shoots and root surfaces ([Fig. 6](#)). Previous studies with rice also reported  
311 higher content of As in rice roots ([Abedin et al., 2002](#)). The higher storage of As in roots and  
312 lower translocation to shoots can be explained by the reduction of arsenate to arsenite in roots,  
313 complexation with thiols, and sequestration in the root vacuoles ([Zhao et al., 2009](#)).

314 Formation of Fe-plaque on rice root surfaces and its effect on As uptake in rice have been  
315 well explained in literature ([Liu et al., 2006](#)). Although Fe-plaque inhibits the As uptake ([Zhang](#)  
316 [et al., 1998](#)), increase of the uptake of toxic and nutrient elements in plants and organisms by Fe-  
317 plaque has also been reported ([Ye et al., 2001](#)). The effects of Fe-plaque on the uptake of nutrient

318 and/or toxic elements depend on the amount of Fe-plaque on root surfaces (Zhang et al., 1998).  
319 Otte et al. (1989) reported higher concentration of Zn in roots of *Aster tripolium* L. coated with  
320 500-2000 nmol Fe cm<sup>-2</sup> compared to those coated with less than 500 or more than 2000 nmol Fe  
321 cm<sup>-2</sup>. Even though the increasing amount of Fe-plaque elevates As accumulation on the root  
322 surfaces, it does not affect As uptake in rice shoots. The Fe-plaque acts as “buffer” to prevent the  
323 translocation of As from roots to shoots Liu et al. (2004).

324 Present study also report that the As contents in roots and shoots were higher in rice  
325 seedlings grown with chelating ligands compared to those grown without chelating ligands (Fig.  
326 6). Arsenic content in roots was highest when the rice seedlings were grown with HIDS while  
327 the content was identical when grown with EDTA, EDDS, or IDS. The results suggest that  
328 chelating ligands increased As uptake in rice root significantly, though its translocation form root  
329 to shoot was not increased. The use of chelating ligands, especially the EDTA, EDDS, IDS, etc.  
330 for the increase of heavy metals have been studied extensively (Jean et al., 2008; Marques et al.,  
331 2008). Present study reports a better/comparable performance of HIDS to that of others studied  
332 previously for the first time.

333 Arsenate has a high adsorptive affinity to Fe oxides (Zhao et al., 2009). Chelating ligands  
334 solubilization/desorption As from the Fe-plaque of rice roots, and rice plant readily uptakes  
335 desorbed/soluble As from the nutrient solution. The results of the present study reveal that the  
336 HIDS is stronger then EDTA, EDDS, or IDS for dissolution/desorption of precipitated As. Since  
337 the EDTA is not readily biodegradable, and is persistent in the environment, the biodegradable  
338 HIDS would be a good alternative to EDTA in the phytoextraction/phytoremediation of As.

339

#### 340 **4. Conclusions**

341 The use of chelating ligands in the phytoextraction of toxic metals and in the increase of  
342 essential nutrient elements is not new at all. Especially, the EDTA and EDDS have been widely  
343 used in agriculture for long time to serve the above purposes. The use of EDTA, however, has  
344 the disadvantage that it is quite persistent in the environment due to its low biodegradability.  
345 Therefore, looking for biodegradable chelating ligands is an important concern to the researchers.  
346 In this study the effectiveness of HIDS for the increase of Fe bioavailability and As  
347 phytoextraction was investigated, and the results were compared with those of EDTA, EDDS,  
348 and IDS. The Fe limiting condition was induced by increasing the pH of the growth solution.  
349 Results show that the performance of HIDS was more effective than that of other chelating  
350 ligands. HIDS is a newly introduced, biodegradable and environmentally harmonious chelating  
351 ligands with high chelating capability. Therefore, it would be a good alternative to the EDTA.

352

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356

### 357 **References**

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492 Figure Captions:

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494 Fig. 1: Growth of rice (*Oryza sativa* L.) affected by the pH of nutrient solution. Results are  
495 presented as mean and the error bars express  $\pm$  SD ( $n = 3$ ).

496

497 Fig. 2: Fe concentration in roots, root surfaces, and shoots of rice seedling (*Oryza sativa* L.) as  
498 affected by the pH of nutrient solution. Results are presented as mean and the error bars  
499 express  $\pm$  SD ( $n = 3$ ).

500

501 Fig. 3: Fe uptake and translocation in rice seedling (*Oryza sativa* L.) as affected by chelating  
502 ligand concentrations in the nutrient solution. Results are presented as mean and the error  
503 bars express  $\pm$  SD ( $n = 3$ ).

504

505 Fig. 4: Growth of rice seedling (*Oryza sativa* L.) affected by chelating ligand concentrations in  
506 the nutrient solution. Results are presented as mean and the error bars express  $\pm$  SD ( $n = 3$ ).

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508 Fig. 5: Fe concentration in root surfaces and plant tissues (roots and shoots) of rice seedling  
509 (*Oryza sativa* L.) as affected by chelating ligands and arsenic in the nutrient solution.  
510 As(+) and As(-) indicate with and without arsenic, respectively.

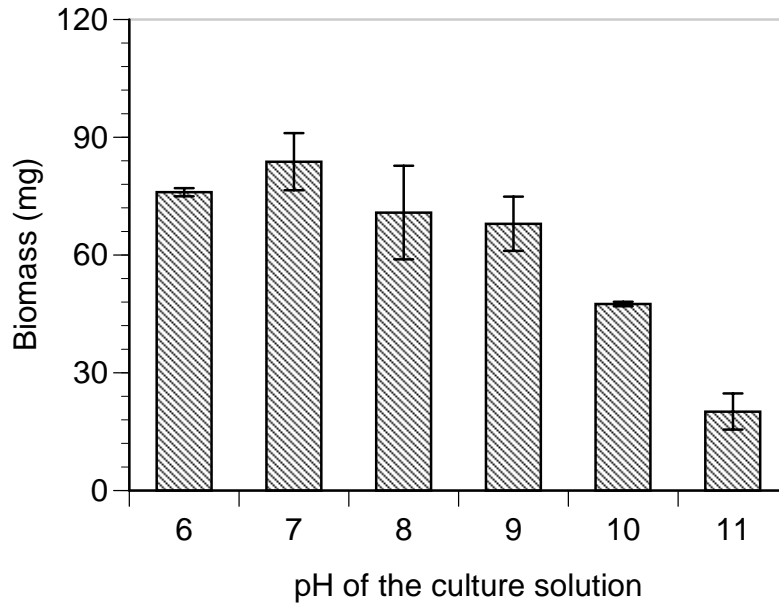
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512 Fig. 6: As concentration in roots, shoots, and root surfaces of rice seedling (*Oryza sativa* L.) as  
513 affected by chelating ligands in the nutrient solution.

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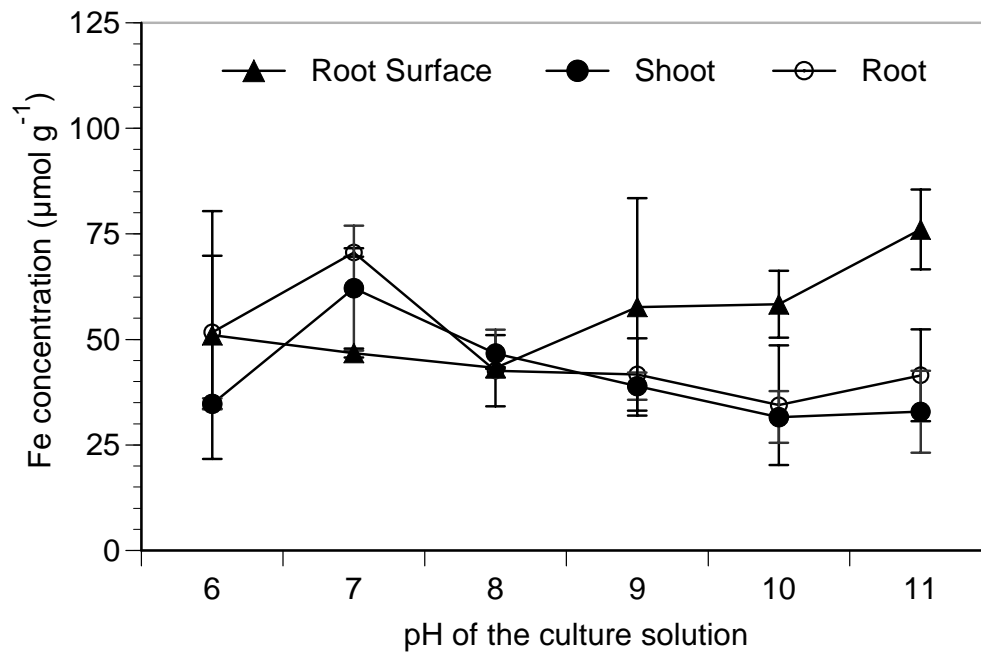
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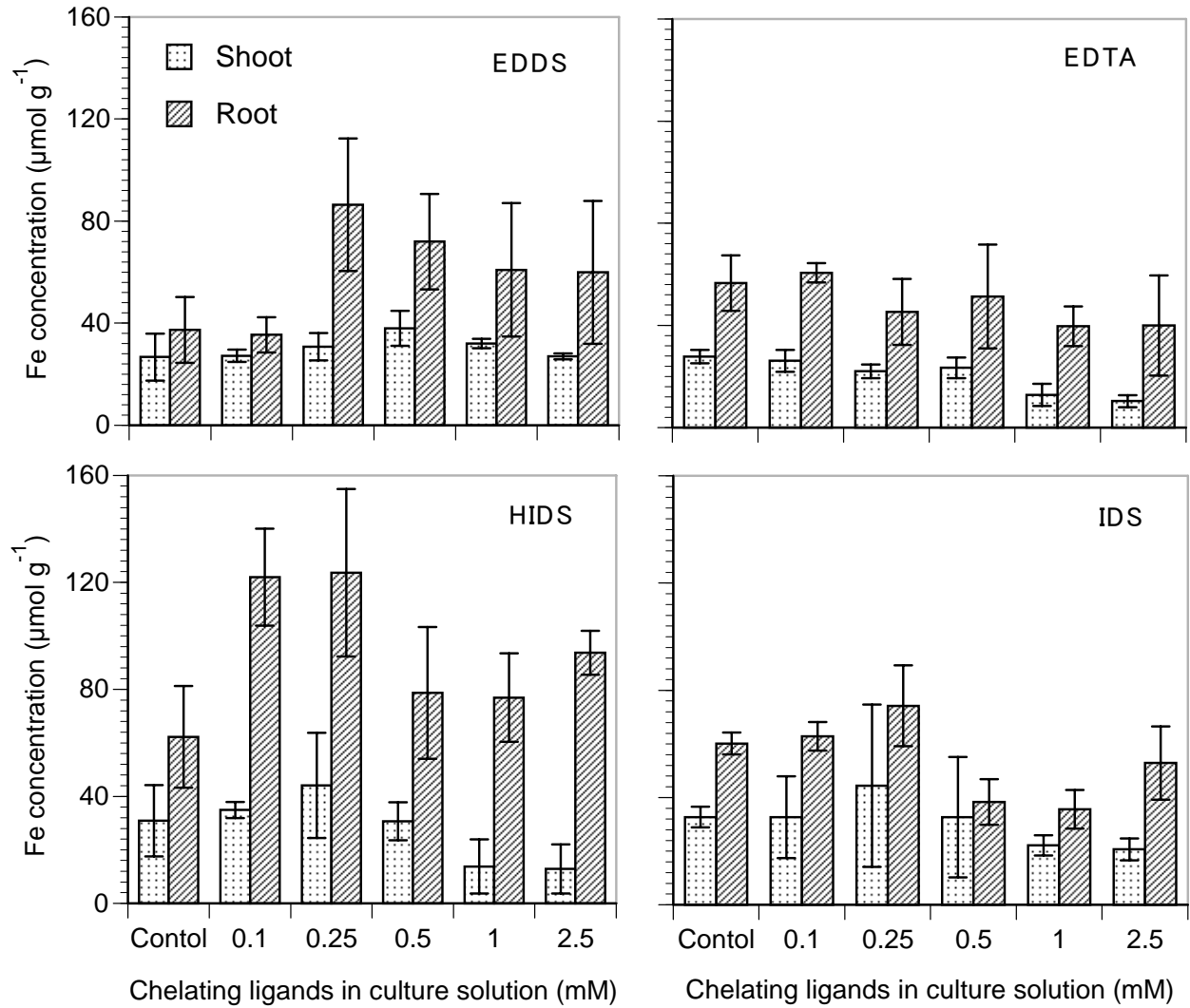
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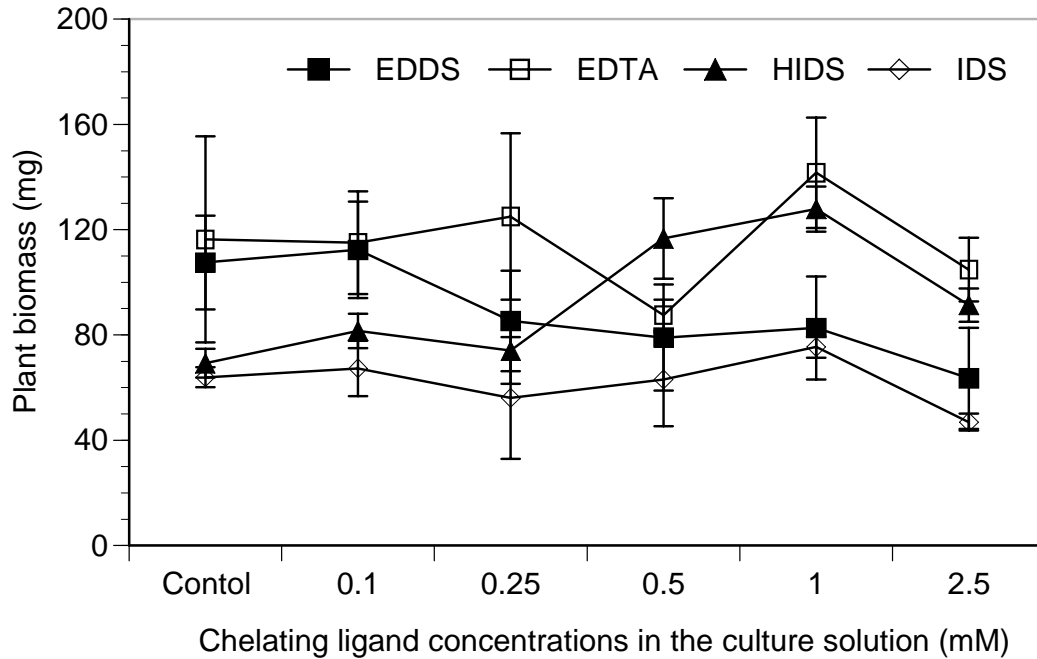
532 bars express  $\pm$  SD ( $n = 3$ ).

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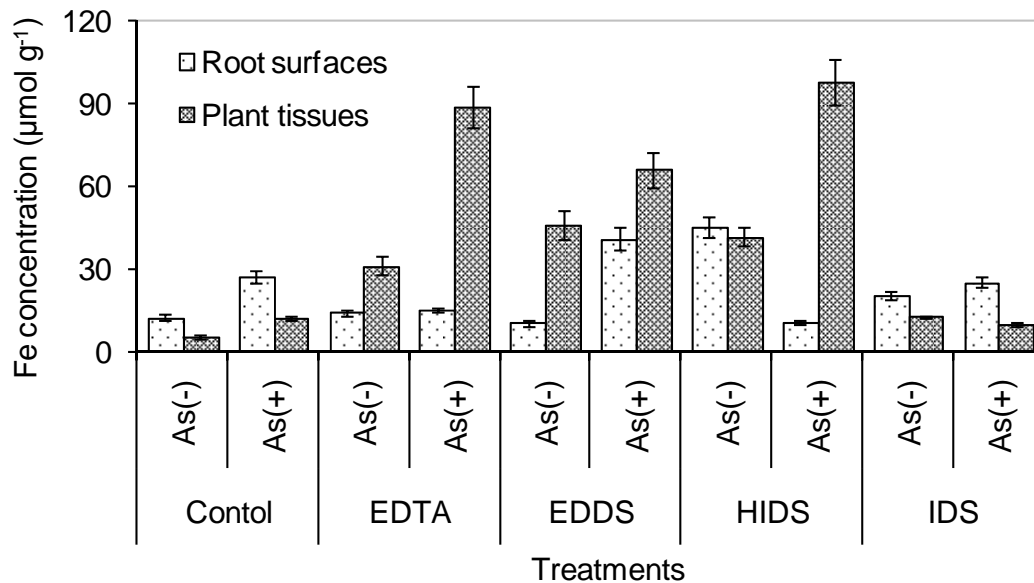


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538 Fig. 4: Growth of rice seedling (*Oryza sativa* L.) affected by chelating ligand concentrations in  
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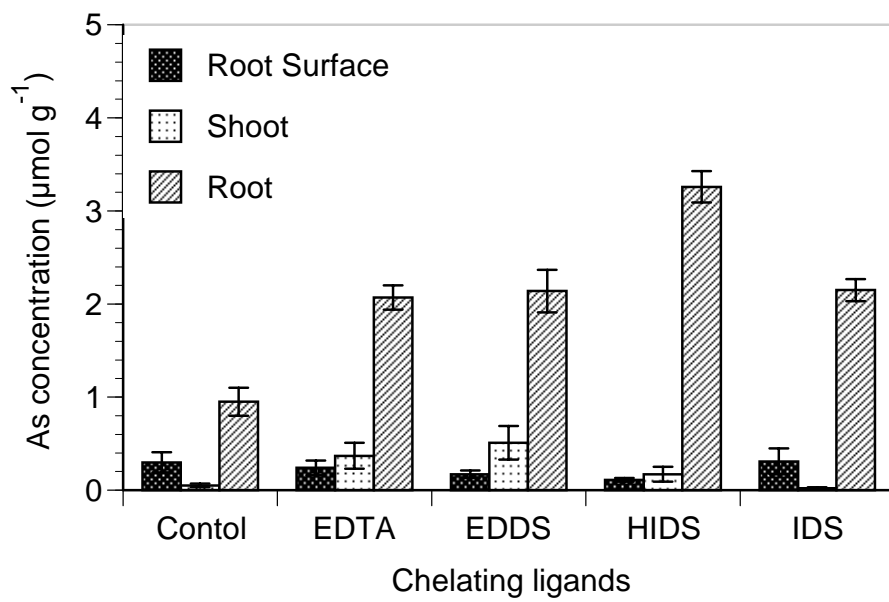
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