Seventeen-Year Observation on Urinary Cadmium and β 2-Microglobulin in Inhabitants After Cessation of Cadmium-Exposure in Japan

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13	Abstract The purpose of this study is to clarify the change and relationship of urinary cadmium
14	(Cd) and β_2 -microglobulin (β_2 -MG) concentrations of inhabitants in Cd-polluted areas after soil
15	restoration. The urinary Cd and β_2 -MG concentrations of 25 males and 28 females did not show a
16	significant change, 22 years after the Cd-polluted soil was restored. Once exposed to Cd, it was
17	found to remain in the body, 22 years after the Cd -polluted soil was restored. However, this did not
18	influence renal tubular dysfunction in most of the younger generation compared with elders heavily
19	exposed to Cd.
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21	Keywords Urinary-Cadmium, Urinary- β_2 -Microglobulin, Biological half life, Soil restoration
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26	Cadmium (Cd), atomic No 48, atomic weight 112 is a mineral found in mines. Cd is a by-product
27	that is released to the environment when zinc is refined from the mines (Wada 1986). Residents in
28	mining areas have been exposed to chronic Cd poisoning for long periods of time, for example the
29	Kakehashi river basin (Ishikawa Prefecture) population were exposed for about 100 years.
30	Kakehashi river basin inhabitants use the river water that Cd is released into for farming vegetables,
31	rice etc. Hence, people were directly exposed by consuming farming foods as well as fish from the
32	river. This oral exposure resulted in kidney dysfunction, osteomalacia, osteoporosis, liver injury and
33	hypertension (Kasuya. 1985; Saito et al. 1993). The most common characteristic of Cd exposure is
34	renal tubular dysfunction. An indicator of renal tubular dysfunction is urinary β_2 -microglobulin
35	(β_2 -MG), a low molecular weight protein which is used as an indicator and judged by the limit 1000
36	$\mu g/gCr$ (Nogawa et al. 1983; Aoshima et al. 1988; Kubota et al. 1985). Due to Cd being released to
37	the Jinzu river (Toyama Prefecture, Japan), inhabitants in the surrounding areas have been exposed
38	for a long time. Itai-itai disease is the severest disease cause by chronic Cd poisoning. The direct
39	translation of the Japanese phrase "Itai-itai" is based on the patient's response "Ouch-ouch" to the
40	many fractures in their bodies by osteomalacia (Aoshima, Kasuya 1993). The ministry of Health in
41	Japan and Welfare recognized Itai-itai disease as the first mining-related illness in April 1968

42 (Matsunami 2006). There are six Cd-polluted areas in Japan. The largest Cd polluted area is Jinzu

43 river basin, followed by Kakehashi river basin (Nakagawa, Nishijo 1999).

In 1981 the Ishikawa Prefecture Government decided to restore the soil in the most Cd-polluted 44 Kanehira areas around Kakehashi river basin (Ishikawa Prefecture 1975) and performed a physical 45examination on Kakehashi river basin inhabitants of over 50 years of age (Ishikawa Prefecture 46South Kaga Public Health 1997). It was evident from the results obtained that the biological 47half-life of Cd is 10-30 years (Friberg 1985) and Cd concentration increased with age (Aoshima, 48Kasuya 1993). The results showed that 14.3% men and 18.7% women were over the 1000 µg/g.Cr 49limit of urinary β_2 -MG. However, only 6% men and 5% women were over the limit in the control 50area. This shows a significant difference (Kido 1995). In 1986, five years after soil restoration 51(Kido et al. 2002) were apprehensive about the effects of Cd on inhabitants 50 years and under, who 52were not examined in 1981 by Ishikawa Prefecture Government. Further tests were performed 4 53times over a 17 year period from 1986 to 2003, in inhabitants of most Cd -polluted area of 54Kakehashi River basin. The purpose of this study is to clarify the change and relationship of urinary 55Cd and β_2 -MG concentrations of inhabitants in Cd- polluted areas after soil restoration. 56

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58 Materials and Method

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In 1981 the Ishikawa Prefecture Government restored the soil in the most Cd-polluted area in the Kakehashi river basin. Five years after soil restoration in 1986, investigations were carried out to determine the influence of Cd on the inhabitant health, and 3 further tests were carried out in 1991, 1999 and 2003, over 17 years. The subjects were 53 inhabitants who received physical examinations 4 times during 17 years. The 53 inhabitants (50 years and under in 1986) included 25 men and 28 women. The mean age of the subjects was 40.2 years old (men: 38.5, women: 42.1) in 1986.

The research process was spread into two parts; 1) Urinary test 2) Questionnaire. The urinary test involved measuring concentration of urinary creatinine (Cr), Cd and β_2 -MG in early morning urine. The concentration of urinary indicators was then corrected by Cr. Urinary β_2 -MG, Cd and Cr were measured by Radioimmunoassay (RIA), Flameless Atomic Absorption Spectrophotometry (Honda et al 1989) and Jaffé methods respectively. The questionnaire involved obtaining basic data from subjects by gender, age, period of residence and Brinkman index.

Urinary Cd and β_2 -MG concentrations were converted into a logarithm scale. The results were statistically analyzed using the one-way layout ANOVA, t-test, Pearson correlation coefficient, and multiple regression analysis. The significant difference was p≤0.05 The software used for analysis was SPSS12.OJ for Windows.

Subject participation in this research was voluntary. An agreement was signed regarding this
matter. The examination performed in 1999 and 2003 were accepted by the Kanazawa Medical
University Bioethical committee. To protect privacy of the individual, only numbers (no names)
were used during examination.

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84 **Results and Discussion**

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Table 1 shows that there is no significant difference in urinary Cd concentration (µg/g.Cr) for men 86 and women during a 17 year period, even though both of them showed decreasing tendency. It can 87 88 be seen from table 2 that there is significant difference (p=0.000) in men during 1991-1999 and 1999-2003. However, in women there is no significant difference during a 17 year period. It can be 89 seen from table 3 that there is a significant difference in 1986 (p=0.000), 1991 (p=0.002), 1999 90 (p=0.003) and 2003 (p=0.000), when comparing the urinary Cd concentration between men and 9192 women. On the other hand, urinary β_2 -MG concentration shows a significant difference between men and women in 1986 (p=0.000), 1999 (p=0.022) and 2003 (p=0.003), but there is no significant 93 difference in 1991 (p=0.174). In both cases it can be clearly seen that the geometric mean is higher 94in females than males, except for mean urinary β_2 -MG concentration in 1991. Table 4 shows that in 9596 25 males there is a significant correlation with urinary Cd and β_2 -MG concentrations in 1999 (p=0.044). On the other hand, in 28 females there is a significant correlation between urinary Cd 97 and β_2 -MG concentration in 1986 (p=0.010) and 2003 (p=0.033). Urinary Cd and β_2 -MG 98 concentrations showed significant relationship in 2003. Fig 1 shows that only two subjects were 99 found to exceed the critical urinary β_2 -MG limit (1000 µg/g.Cr) in 2003. Table 5 shows that there is 100no significant regression between urinary β_2 -MG and urinary Cd, age, period of residence, 101 Brinkman index in 25 males and 28 females, except for significant association in females between 102103 urinary β_2 -MG and age (p=0.036) as well as Brinkman index (p=0.040) in 2003.

As shown in table 1 there is no significant difference in urinary Cd concentration in 25 males and 28 females, 22 years after cessation of Cd exposure (soil restoration in 1981, first investigation 5 years after in 1986 until 2003). This shows that once exposed to Cd it remains in the body for a long time. This agrees with the biological half life of Cd which is 10-30 years (Nordberg 1985). This shows that once exposed to Cd, irreversible injury (renal tubular dysfunction) was noted to occur when urinary β_2 -MG excretion exceeded 1000 µg/g.Cr (Cai 2001; Iwata et al. 1993; Kido et al. 1988).

As shown in table 2 in males there is no significant difference in urinary β_2 -MG concentration 111 112over 17 years. However, between 1991-1999 and 1999-2003 (p=0.000) there is a significant 113 decrease and increase respectively. At the moment there is no plausible reason for this "V" shaped curve. But the geometric means of urinary β_2 -MG concentration during 17 years were within the 114normal range. This can be explained by the physiological changes of the human body. Urinary 115 β_2 -MG is a sensitive indicator of renal tubular dysfunction (Nogawa et al. 1983; Aoshima et al. 1161988; Kubota et al. 1985). Previous studies have shown that there is a significant relationship 117118between urinary Cd and β_2 -MG (Kido et al. 1988; Honda et al. 1982). It can be seen from table 3 119that females have a higher geometric mean urinary Cd and β_2 -MG concentration relative to males (Kido et al. 2001). This study shows that urinary Cd still remain in the body 22 years after cessation 120of exposure. Moreover, urinary β_2 -MG remains in the body for the same period of time as Cd 121122(Tables 1 and 2). Therefore, there is a significant correlation between urinary Cd and β_2 -MG 123concentration in males in 1999 (p=0.004) and in females in 1986 (p=0.010) and 2003 (p=0.033) (Table 4). In multiple regression analysis with urinary β_2 -MG, urinary Cd, age, period of residence 124and Brinkman index are contributing factors for Cd exposed subjects. The results in table 5 show 125

that in 2003 there is a significant relationship between urinary β_2 -MG and age (p=0.036) and also Brinkman index (p=0.040) in females. Even though Cd remained in the body for 22 years after cessation of Cd exposure it did not influence renal tubular dysfunction. This is because when the first investigation was conducted in 1986, 5 years after cessation of Cd exposure, the ages of all 53 subjects were under 50 years. At this time their exposure to Cd was less than that of inhabitants 50 years and over. Only four subjects were found to exceed the critical urinary β_2 -MG limit (1000 µg/g.Cr) during 17 years, as shown in figure 1.

Recently, not many long term epidemiological studies have been performed internationally. So it 133is imperative that long term epidemiological studies should be carried out such as our investigation. 134This study has thoroughly examined urinary Cd and β_2 -MG in inhabitants after cessation of Cd 135exposure. In Cd-polluted areas of Japan, this is the first report that shows inhabitants having no 136adverse health effects. Countermeasure of replacement of Cd-polluted soil might contribute to 137younger inhabitants rather than the elder people in the former Cd-polluted areas. Even though the 138seventeen year observation data shows no significance in relationships between urinary Cd and 139140 β_2 -MG due to the aforementioned reasons, further investigations will be carried out to confirm these present results. 141

142 This manuscript can be concluded by the following points:

- i) The change of urinary Cd and β_2 -MG concentrations in 25 males and 28 females were not significant during 17 years after the Cd-polluted soil was restored even though they showed decreasing tendency.
- 146 ii) The comparison with urinary Cd and β_2 -MG concentrations between men and women 147 showed a significant difference during a 17 year period (1986, 1991, 1999 and 2003), except 148 for urinary β_2 -MG concentrations in 1991. The geometric mean was higher in females than 149 males.
- iii) Once exposed to Cd, it was found to remain in the body, 22 years after the Cd-polluted soil
 was restored. However, this did not influence renal tubular dysfunction generally..
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169**Table 1** Change in urinary Cd concentration ($\mu g/g.Cr$) in males and females

Iubic I	(one- way layo	-		ion (µg/g.cr) n	i mares and ren	liulos		
	Males (N=25)			Female (N=28)				
Ye	ar	GM	GSD		GM	GSD		
19	86	3.02	2.14		5.50	2.19		
19	91	2.75	2.00		5.50	2.30		
19	99	2.19	1.66		3.80	2.09		
20	03	1.95	1.74		4.27	2.00		
GM: Geo	ometric Mean	GSD: Geometric Standard						
Table 2	Change in uri (one-way lay	• •		entration (μg/g.	Cr) in males ar	nd female	es	
		Mal	les (N=2	5)	Females	s (N=28)		
Ye	ar	GM	GSD		GM	GSD		
19	86	30.20	8.32		83.18	4.07		
19	91	89.13	2.51	p=0.000	128.83	2.95		
1	999	12.88	6.46	p=0.000	51.29	10.23		
2003		57.54	2.24	p=0.000	131.83	3.02		
GM: Geo	ometric Mean		GSD:	Geometric Star	ndard Deviation p: p-value			
Table 3	Comparison w 28 females (t-te	est)			ntration (μg/g.C			
Veen	Caralan		•	(µg/g.Cr)		$ry \beta_2 - MO$		
Year	Gender	GM	GSD	p-value	GM 20.00	GSD	p-v	
1986	M	3.05	2.16	p=0.000	29.90 82.12	8.36	p=0.0	
	F	5.52 2.70	2.19		83.12	4.09		
1991	M	2.79	2.01	p=0.002	88.19	2.47	p=0	
	F	5.47	2.27		128.94	2.94		
1999	Μ	2.17	1.66	p=0.003	12.84	6.49	р=(
1///	Ę	2 70	() 1 ()	P 0.002			P-4	
1777	F	3.79	2.10	P 0.000	50.90	10.15	Р-(
2003	F M F	3.79 2.00 4.30	2.10 1.73 2.00	p=0.000	50.90 57.50 133.05	10.15 2.20 3.01	р=(

GM: Geometric Mean 182

GSD: Geometric Standard Deviation

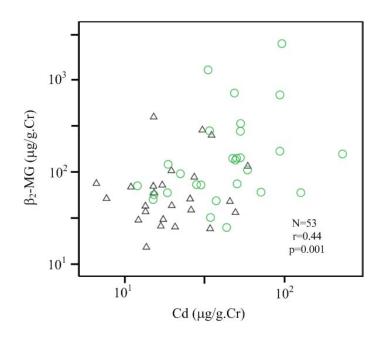




Figure. 1 Scatter plot of correlation between urinary Cd and β_2 -MG concentration for 28 females (\bigcirc) and 25 males (\triangle) in 2003 (Pearson correlation method; r: Pearson correlation coefficient).

Table 4 The correlation coefficient between urinary Cd and β 2-MG concentration

in females and males (Pearson correlation method)					
	1986	1991	1999	2003	
Males	r=0.202	r=0.183	r=0.349	r=0.155	
(N=25)	p=0.167	p=0.191	p=0.044	p=0.229	
Females	r=0.435	r=0.241	r=0.093	r=0.351	
(N=28)	p=0.010	p=0.108	p=0.320	p=0.033	
r : Pearson corre	lation coefficient	p:p-val	lue		

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Table 5 Multiple regression analysis of urinary β_2 -MG and Age, urinary Cd,

					ales and fer	males				
-	Males (N=25)									
-	1986		1991		1999		2003			
		β	p-value	β	p-value	β	p-value	β	p-value	
	Age	-0.142	0.682	-0.037	0.914	-0.174	0.501	-0.073	0.804	
	Cd	0.174	0.593	-0.114	0.775	0.268	0.284	0.038	0.907	
	Period of residence	0.145	0.651	0.241	0.426	0.137	0.632	0.254	0.425	
_	Brinkman index	0.125	0.674	0.444	0.132	0.401	0.119	0.118	0.689	
_	\mathbf{R}^2	R^2 0.055		0.	0.140		0.232		0.056	
_										
_				Females	(N=28)					
_		1	986	1	991	1	999	2003		
		β	p-value	β	p-value	β	p-value	β	p-value	
	Age	0.111	0.684	0.247	0.410	0.301	0.177	0.426	0.036	
	Cd	0.294	0.327	-0.056	0.858	-0.181	0.386	0.015	0.941	
	Period of residence	0.112	0.625	9.225	0.340	0.397	0.053	0.244	0.208	
	Brinkman index	-0.037	0.847	-0.083	0.676	-0.211	0.237	-0.346	0.040	
_	\mathbb{R}^2	0.	435	0.134		0.322		0.435		
	Q. Standardinad name	-1		• ,		R ² : Coefficient of determination			4.00	
	β: Standardized parti	al regressi	ion coeffic	lent		R : Coe	incient of	determina	tion	
	p: Standardized parti	al regressi	ion coeffic	ient		R : Coe	incient of	determina	uon	
	p: Standardized parti	al regress	ion coeffic	eient		K : Coe	incient of	determina	uon	
	p: Standardized parti	ai regress		ient		K : Coe	incient of	determina	uon	
	p: Standardized parti	al regress:		ient		K : Coe	incient of	determina	uon	
	p: Standardized parti	al regress	ion coeffic	ient		K : Coe	incient of	determina	tion	
	p: Standardized parti	al regress	ion coeffic	ient		K : Coe	incient of	determina	tion	
	p: Standardized parti	al regress	ion coeffic	ient		K : Coe	incient of	determina	tion	
	p: Standardized parti	al regress	ion coeffic	ient		K : Coe	incient of	determina	tion	
	p: Standardized parti	ai regress	ion coeffic	ient		R : Coe	incient of	determina	tion	
	p: Standardized parti	al regress:	ion coeffic	ient		K : Coe	incient of	determina	tion	
	p: Standardized parti	al regress:	ion coeffic	hent		R : Coe	incient of	determina	tion	
	p: Standardized parti	al regress:	ion coeffic	hent		R : Coe	incient of	determina	tion	
	p: Standardized parti	ai regress	ion coeffic	hent		R : Coe	incient of	determina	tion	
	p: Standardized parti	ai regress	ion coeffic	hent		R : Coe	ancient of	determina	tion	
	p: Standardized parti	ai regress	ion coeffic	hent		R : Coe	ancient of	determina	tion	
	p: Standardized parti	ai regressi	ion coeffic	hent		R : Coe	ancient of	determina	tion	
	p: Standardized parti	ai regress	ion coeffic	hent		R : Coe	ancient of	determina	tion	

period of residence, Brinkman index in males and females

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