

Hydroxyiminodisuccinic acid (HIDS): A novel biodegradable chelating ligand for the increase of iron bioavailability and arsenic phytoextraction

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24 Abstract

25 The influence of biodegradable chelating ligands on arsenic and iron uptake by hydroponically
26 grown rice seedlings (*Oryza sativa* L.) was investigated. Even though the growth solution
27 contained sufficient Fe, the growth of rice seedlings gradually decreased up to 76% with the
28 increase of pH of the solution from 7 to 11. Iron forms insoluble ferric hydroxide complexes at
29 neutral or alkaline pH in oxic condition. Chelating ligands produce soluble 'Fe-ligand complex'
30 which assist Fe uptake in plants. The biodegradable chelating ligand hydroxyiminodisuccinic
31 acid (HIDS) was more efficient than those of ethylenediaminetetraacetic acid (EDTA),
32 ethylenediaminedisuccinic acid (EDDS), and iminodisuccinic acid (IDS) in the increase of Fe
33 uptake and growth of rice seedling. A total of 79 ± 20 , 87 ± 6 , 116 ± 15 , and 63 ± 18 mg dry biomass
34 of rice seedlings were produced with the addition of 0.5 mM of EDDS, EDTA, HIDS, and IDS
35 in the nutrient solution, respectively. The Fe concentrations in rice tissues were 117 ± 15 , 82 ± 8 ,
36 167 ± 25 , and 118 ± 22 $\mu\text{mol g}^{-1}$ dry weights when 0.25 mM of EDDS, EDTA, HIDS, and IDS
37 were added to the nutrient solution, respectively. Most of the Fe accumulated in rice tissues was
38 stored in roots after the addition of chelating ligands in the solution. The results indicate that the
39 HIDS would be a potential alternative to environmentally persistent EDTA for the increase of Fe
40 uptake and plant growth. The HIDS also increased As uptake in rice root though its translocation
41 from root to shoot was not augmented. This study reports HIDS for the first time as a promising
42 chelating ligand for the enhancement of Fe bioavailability and As phytoextraction.

43

44

45 **Keywords:** Arsenic, Iron, Chelating ligands, Rice (*Oryza sativa* L.), Hydroponics, Bioavailable,

46 HIDS.

47

48 **1. Introduction**

49 Iron is an essential micronutrient for plants, which plays important roles in respiration,
50 photosynthesis, and many other cellular functions such as DNA synthesis, nitrogen fixation, and
51 hormone production (Vert et al., 2002). Although abundant in nature it forms insoluble ferric
52 hydroxide complexes (also known as Fe-plaque) at neutral or alkaline pH in oxic condition
53 (Guerinot and Yi, 1994). The Fe-plaque formation in the rhizosphere soils, however, results in
54 the Fe deficiency to plants. In nature, rhizospheric microbes exude siderophores to the root-
55 plaque interface. These siderophores solubilize ferric iron in the rhizosphere, render its
56 bioavailability, and plants uptake the Fe by specific membrane receptors (Romheld, 1987).

57 Elevated levels of As in soil from natural and anthropogenic sources is a threat to plants'
58 health (Rahman et al., 2008). Remediation of contaminated soil is important to prevent As
59 deposition in food crops and its subsequent transfer into the human body through the food chains
60 (Rahman et al., 2008). Phytoremediation becomes a promising alternative and environmentally
61 safe technology for the remediation of environmental pollutants (Raskin et al., 1997; Tu et al.,
62 2002). An essential prerequisite for phytoremediation of contaminated soil is solubility and
63 bioavailability of As (Fitz and Wenzel, 2002). But the solubility and bioavailability of As
64 becomes reduced by adsorption to variable charged minerals (Fe and Al) at alkaline pH (Xu et al.,
65 2008). In the past decade, chelant-enhanced phytoremediation has received much attention
66 (Pastor et al., 2007). This technique aims to cleanse polluted soils by solubilizing the toxic
67 metals, allowing it to be accumulated in plants that would subsequently remove toxic metal from
68 the site. Publications on chelant-enhanced phytoremediation have increased steadily to about 15-
69 20 per year in the last few years, indicating that this is a growing and active research field
70 (Nowack et al., 2006).

71 Research on the interaction of plants with chelating ligands started in the 1950s with a
72 view to reduce the deficiencies of the essential nutrients such as Fe, Mn, Cu, and Zn (Wenger et
73 al., 2005). Among all soil-applied Fe fertilizers, synthetic Fe(III)-chelates, mainly Fe(III)-
74 chelates of polyaminecarboxylic acids with phenolic groups, such as ethylenediamine di(*o*-
75 hydroxyphenylacetic) acid (EDDHA), and ethylenediamine di(2-hydroxy-4-methylphenylacetic)
76 acid, are the most effective and commonly used (Alvarez-Fernandez et al., 2005). On the other
77 hand reports on As phytoextraction by chelating ligands is limited though a number of
78 investigations have been conducted on chelant-enhanced phytoextraction of Pb, Zn, Hg, Cu and
79 some other heavy metals (Luo et al., 2005). Ethylenediaminetetraacetic acid (EDTA) has been
80 very popular to achieve this purpose, but it is quite persistent in the environment because of its
81 low biodegradability. This, in combination with its high affinity for heavy metal complexation,
82 results in an increased risk of leaching. EDTA also impairs plant growth severely, even at low
83 concentrations (Bucheli-Witschel and Egli, 2001).

84 Biodegradable chelating ligands, such as ethylenediaminedisuccinic acid (EDDS),
85 Hydroxyiminodisuccinic acid (HIDS), and iminodisuccinic acid (IDS) would be good choice and
86 alternative to less biodegradable EDTA. The physicochemical properties of EDDS, EDTA, and
87 IDS have already been discussed and tested for the phytoextraction of heavy metals by a number
88 of researchers (Helena et al., 2003; Evangelou et al., 2007). HIDS is a new chelating ligand
89 introduced by Nippon Shokubai Co. Ltd. It is one of the highly biodegradable (biodegradation
90 rate is about 22.4% within 48 h) and safe chelating ligands. It traps and inactivates various kinds
91 of metals ions over a wide range of pH, particularly Fe^{3+} and Cu^{2+} , as well as Ca^{2+} and Mg^{2+} ;
92 shows high stability in harsh conditions and high temperature (80 °C); is highly soluble in
93 aqueous alkaline solution (Sokubai, 2009). Because of high degradation rate and high stability
94 constant with Fe^{3+} ($\text{pK}_a\text{Fe}^{3+}$ is 12.5) of HIDS, we become interested to investigate the
95 effectiveness of the chelating ligand for the increase of Fe bioavailability and phytoremediation

96 of As. The EDTA, EDDS, and IDS were also used in the present study to compare the results of
97 HIDS. Our research approach was to find a biodegradable and eco-friendly chelating ligand that
98 is more desirable than EDTA or EDDS for Fe bioavailability and As phytoextraction.

99

100 **2. Materials and Methods**

101 *2.1. Seed sterilization*

102 Rice seeds of BRRI dhan 29 were collected from Bangladesh Rice Research Institute.
103 The seeds were surface-sterilized before using them in the experiment. For sterilization, about
104 100 g seeds were soaked in 200 mL of 1% methyl-1-butylcarbamoyl-2-benzimidazole carbonate
105 solution for 10 min. After that, the seeds were washed by deionized (DI) water (using an E-pure
106 system (Barnstead)) and kept in DI water at 20 °C for 24 h. The seeds were then washed and
107 transferred to DI water of 45 °C for 2 min, and of 52 °C for 10 min.

108

109 *2.2. Chemicals*

110 Stock solutions of EDTA, EDDS, HIDS, and IDS were prepared by dissolving
111 ethylenediamine-N,N,N',N'-tetraacetic acid (Dojindo Molecular Technologies, Japan),
112 ethylenediamine-N, N'-disuccinic acid (Chelest), tetrasodium 3-hydroxy-2,2'-iminodisuccinate
113 (Nippon Syokubai, Japan), and tetrasodium iminodisuccinate (Bayer) in 0.1 M sodium hydroxide,
114 respectively. Other reagents were of analytical grade or better. All solutions were prepared with
115 DI water.

116

117 *2.3. Nutrient solution*

118 Sterilized rice seeds were germinated on pre-sterilized bloating paper (seed bed) with
119 standard murashige and skoog (MS)([Murashige and Skoog, 1962](#)). Iron concentration in the

120 experimental solution was 0.36 mM while its concentration was 27.8 mg L⁻¹ in pre-experimental
121 solution (used for growing rice seedling prior to the experiment). The pH of the pre-experimental
122 solution was adjusted to 6.5 while the pH of experimental solution was 9.0. Rice seedlings were
123 grown on the seed bed for 1 wk. In preparing MS culture solution, FeSO₄·7H₂O was used as Fe
124 source instead of NaFe(III)-EDTA.

125

126 *2.4. Experimental setup*

127 Rice seedlings were transferred to the experimental solution after one week of growth in
128 pre-experimental solution. In the experimental solution, rice seedlings were grown in two steps.
129 In the first step, rice seedlings were grown with different concentrations of chelating ligands (up
130 to 2.50 mM) to observe the effect of chelating ligands on Fe uptake. In the second step, 6.0 μM
131 of As (Na₂HAsO₄·7H₂O) was added to the nutrient solutions containing 1.0 mM of chelating
132 ligands to see the effect of chelating ligands on Fe and As uptake. Iron concentration in the
133 experimental solution was 0.36 mM, and the pH of the solution was adjusted to 9 using 0.1 M
134 KOH. About 100 mL of the solution was taken into 250-mL polystyrene bottles with three
135 replications, and three uniform seedlings were cultivated in each bottle. The experiment was
136 performed following randomized design. Rice plants were grown in a plant growth chamber and
137 the conditions in the chamber were set as 14:10 h light/dark schedule, 100-125 μ E m⁻² s⁻¹ light
138 intensity, 22(±2) °C temperatures. Rice seedlings were grown in experimental solution for 5 d.

139

140 *2.5. CBE-extraction of Fe-plaques*

141 At harvest, the shoots were cut from 1 cm above the roots and separated. The Fe-plaques
142 from root surfaces were extracted using citrate-bicarbonate-ethylenediaminetetraacetate (CBE)-
143 technique, a modified method of dithionite-citrate-bicarbonate extraction by [Taylor and Crowder](#)
144 [\(1983\)](#) to determine the real amount of Fe and As contents in rice tissues. The CBE solution was

145 prepared from 0.03, 0.125 and 0.050 M of sodium citrate, sodium bicarbonate, and EDTA,
146 respectively. Roots were treated with 30 mL of CBE solution for 60 min at room temperature.
147 The roots were then rinsed with deionized water for 3 times, and the rinsed water was added to
148 the CBE-extracts to make a total of 30 mL.

149

150 *2.6. Sample preparation*

151 After rinsing with deionized water for four times, the root samples were kept on clean
152 absorbent paper to remove the water from the root surfaces. Both the root and the shoot samples
153 were dried at 65 °C until they reached in a constant weight. Then the dried samples were
154 weighted and taken into 50-mL polyethylene tubes for digestion. Five mL of 65% HNO₃ were
155 added to the sample and kept for 12 h. The samples were heated on a heating block at 95 °C for 2
156 h. After cooling to room temperature, 3 mL of 30% hydrogen peroxide were added, and the
157 samples were heated again at 105 °C for 20 min. Then, the digests were diluted to 30 mL with DI
158 and analyzed for As and Fe.

159

160 *2.7. Chemical analysis*

161 Arsenic and Fe were analyzed using graphite-furnace atomic absorption spectrometer (Z-
162 8100, Hitachi, Japan). Certified standard reference material 1573a (tomato leaf from NIST,
163 USA) was used to check the accuracy of analysis. Arsenic concentration in certified standard
164 reference materials was $0.112 \pm 0.004 \mu\text{g g}^{-1}$ dry weight (all the reported data in this article are
165 expressed as dry weight) while the measured concentration was $0.114 \pm 0.002 \mu\text{g g}^{-1}$. The
166 concentrations detected in all samples were above the instrumental limits of detection (≥ 0.01
167 μM in water sample).

168 All chemical reagents used in this experiment were of analytical grade. Glassware and
169 dishes were washed with detergent and 1 N HCL solution, and rinsed with DI water for eight

170 times before use. In each analytical batch, at least two reagent blanks and three replicate samples
171 were included.

172

173 **3. Results and Discussions**

174 *3.1. Effect of pH on rice growth*

175 Rice seedlings were grown in nutrient solution adjusted to different pH ranging between
176 6 and 11. Results show that the biomass production of rice seedlings was affected by the pH
177 significantly. The highest biomass of rice seedling (83 ± 7 mg) was observed at pH 7, which was
178 about 16, 19, 43, and 76% higher than those at pH 8, 9, 10, and 11, respectively (Fig. 1). The rice
179 growth remain unchanged, and even died at pH 10 and 11. Rice plants have a tendency of higher
180 Fe uptake than that of other plants (Becker and Asch, 2005). But the pH of the growth medium
181 plays an important role in Fe bioavailability and uptake. Even though the Fe is sufficient in
182 growth medium, it forms insoluble ferric hydroxide complexes at alkaline pH in oxic condition
183 (Cohen et al., 1998). Therefore, Fe bioavailability and uptake decreases drastically. In the present
184 study, it was observed that the Fe concentrations in tissues of rice seedlings were highest at pH 7
185 compared to those at other pHs (Fig. 2). This trend of Fe uptake in rice tissues is correlated to
186 that of biomass production of rice seedlings (Fig. 1). The result implies that the influence of pH
187 on rice growth is the ultimate effect of reduced Fe bioavailability and uptake. Moreover, Fe
188 concentrations on root surfaces of rice seedlings were lowest at pH 7 and 8 compared to those at
189 other pHs (Fig. 2). High level of Fe on root surfaces of rice seedling at pH 11 reveals the
190 formation of Fe-hydroxides (Fe-plaque) on root surfaces, which decreased the Fe uptake in rice
191 tissues. Formation of Fe-plaques on the roots of wetland plants (Hansel et al., 2001) and
192 hydroponically grown rice seedling (Hu et al., 2005) have also been reported. The precipitation
193 of ferric (oxyhydro)-oxides (FeO_x) and its association with phytoplankton surfaces, both in

194 natural conditions and laboratory cultures, has been reported by [Tang and Morel \(2006\)](#).
195 [Robinson et al. \(2006\)](#) also found the occurrence of Fe-plaque on aquatic macrophytes collected
196 from the Taupo Volcanic zone, New Zealand.

197 The Fe deficiency results in Fe-chlorosis in green leaves, which retards plant growth, and
198 leads to the reduction of crop yields ([Guerinot and Yi, 1994](#)). The results of the present study
199 also reveal that the growth of rice seedling decreased drastically at higher pH, which is the
200 consequence of Fe-chlorosis.

201

202 *3.2. Influence of chelating ligands on Fe uptake-translocation*

203 Influence of EDDS, EDTA, HIDS, and IDS on Fe uptake and translocation in rice
204 seedlings were investigated at different concentrations of the ligands ranging between 0.1 and
205 2.5 mM. Results showed that Fe uptake in rice seedling differed significantly with the type and
206 concentrations of the chelating ligands. Iron uptake was highest at 0.25 mM of the chelating
207 ligands compared to the control treatment. Iron uptake decreased gradually with the increase of
208 chelating ligand concentrations above 0.25 mM ([Fig. 3](#)). The effectiveness of HIDS and EDDS
209 in the increase of Fe uptake in rice tissues was higher than that of EDTA and IDS. Iron
210 concentrations in roots of rice seedling were 35 ± 3 and 44 ± 2 $\mu\text{mol g}^{-1}$ when the HIDS
211 concentrations in the nutrient solution were 0.10 and 0.25 mM, respectively. These
212 concentrations were significantly higher than those of other chelating ligands.

213 Iron concentrations in shoots of rice seedlings were significantly lower than those in roots,
214 and were about identical up to 0.25 mM of chelating ligand treatment. Iron content in shoots
215 decreased with the gradual increase of chelating ligands from 0.25 to 2.50 mM ([Fig. 3](#)). The
216 results indicate that the translocation of Fe from roots to shoots was not affected by lower dose
217 of the chelating ligands. The translocation of Fe was inhibited by the chelating ligands at higher
218 doses (> 0.25 mM).

219 Although abundant in nature, Fe is often unavailable to plants, especially at neutral or
220 alkaline pH, because of the formation of insoluble ferric hydroxide complexes in oxic condition
221 (Robinson et al., 2006). Precipitation of Fe in the rhizosphere, however, may result in the Fe
222 deficiency to the plants. Chelating ligands are used in agriculture as additives in micronutrient
223 fertilizers for the increase of Fe bioavailability. Although some chelating ligands have been
224 reported to increase Fe uptake/translocation in plant, inhibition of Fe uptake/translocation by
225 ligands has also been reported. Chaney et al. (1972) reported that
226 bathophenanthrolinedisulfonate (BPDS) was the most effective inhibitor of Fe
227 uptake/translocation, followed by EDTA > DTPA (diethylenetriaminepentaacetic acid) > CDTA
228 (diaminocyclohexanetetraacetic acid) >> EDDHA. The BPDS inhibited ⁵⁹Fe movement to the
229 exudate by 99.7% even at the lowest level of competitor. The BPDS inhibits Fe translocation by
230 10-100 times compared to those of EDTA, DTPA, or CDTA. Chaney et al. (1972) also observed
231 that EDDHA, the chelator with the highest Fe³⁺ stability constant, only slightly inhibited or
232 actually promoted Fe uptake/translocation, whereas the BPDS with the highest Fe²⁺ stability
233 constant was a severe inhibitor. Thus, stability constant of Fe-ligand (logK_{FeL}) would be one of
234 the important determinants for the promotion or inhibition of Fe uptake/translocation.

235 3.3. Effect of chelating ligands on rice growth

236 Rice seedlings were grown in alkaline nutrient solution (pH 9) containing 0.10, 0.25, 0.50,
237 1.00, and 2.50 mM of chelating ligands and 0.36 mM of Fe. Results show that the growth of rice
238 seedlings was increased with the increase of HIDS and EDTA concentrations up to 1.0 mM, and
239 the growth was decreased at 2.5 mM of chelating ligand concentrations (Fig. 4). The highest
240 biomass production (141±21 mg) of rice was observed when 1.0 mM of EDTA was added to the
241 nutrient solution followed by 127±8, 82±19, and 75±4 mg for HIDS, EDDS, and IDS,
242 respectively.

243 Chelating ligands have been used to enhance Fe bioavailability (Alvarez-Fernandez et al.,
244 2005). The concentration of chelating ligands in the nutrient medium is important for the
245 solubilization of precipitated Fe and the increase of its bioavailability. In the present study, it was
246 observed that the rice seedling produce highest biomass at 1.0 mM chelating ligand
247 concentrations, and the growth remain unchanged, and even died at higher concentration (>1.0
248 mM).

249 Although the growth of all organisms is dependent on the acquisition of the proper
250 quantities of trace elements, excess amount of some metals such as Fe, zinc, manganese, and
251 copper produce toxic effects (Morel and Hering, 1993). However, ferric ions and their complexes,
252 which have low solubility in aquatic system, are extensively buffered by chelation (Morel and
253 Hering, 1993), and increase their dissolved concentration. The dissolved concentration of Fe
254 determines its rate of uptake by the organisms. Anderson and Morel (1982) reported that the Fe
255 uptake rate in laboratory cultures of the marine diatom *Thalassosira weissflogii* is a unique
256 function of the free ferric ion concentration at the presence of 10^{-5} M of various chelating ligands
257 (1.4×10^7 cells L^{-1}). Hudson and Morel (1990) reported that in Fe-limited culture of marine
258 diatom *Thalassosira weissflogii* (10^7 cells L^{-1}) containing 10^{-8} M Fe and 10^{-5} M EDTA and with
259 white-light illumination, both the thermal dissociation of FeEDTA and its photoreduction and
260 reoxidation contribute to the formation of the dissolved inorganic Fe(III) pool responsible for the
261 Fe uptake. In this case, growth of rice seedlings was inhibited by the free ferric ion that was
262 increased by the addition of higher level of chelating ligands.

263 Toxicity of chelating ligands on plants has not been studied extensively. So, it is difficult
264 to interpret the direct toxicity of chelating ligands on plants. Since most of the chelating ligands
265 are synthetic compounds, no nutrient carriers in the plasma membrane are thought to exist
266 (Berne and Levy, 1998). Also, synthetic chelates cannot slip through the plasma membrane as
267 they are too large and polar to move through the plasma lemma lipid bilayer (Berne and Levy,

268 1998). Tanton and Crowdy (1972) observed that most solutes moved into some endodermal
269 passage cells adjacent to the casparian strip intracellularly to the other side of the strip, and then
270 extracellularly to the xylem. The passage cells may include the aquaporins and there may be
271 selectivity toward molecules. Paul et al. (2003) reported that Swiss chard uptakes a considerable
272 amount of EDTA from chelator-buffered hydroponic solution through transpirational flow that
273 occurs via apoplastically.

274

275 *3.4. Influence of chelating ligands and As on rice growth*

276 Chelating ligand treated rice seedlings were grown with and without As to investigate the
277 effect of As and chelating ligands on rice growth. Results show that As does not have a
278 consistent effect on rice growth as chelating ligand has. Rice growth was not affected by As
279 when chelating ligand was not treated. The highest growth of rice seedling was observed in
280 HIDS treated medium. The inconsistent effect of chelating ligand and As on rice growth suggest
281 that in the presence of chelating ligands lower level of As in the growth medium does not affect
282 rice growth significantly. It has been reported that rice growth is not affected by low level of As
283 though the growth decrease drastically with the increase of As in the soil. Abedin and Meharg
284 (2002) also reported that low level of As in water (about 2.0 mg L⁻¹) does not show toxicity to
285 both rice germination and rice growth, but the rice germination and growth were adversely
286 affected by higher As level.

287

288 *3.5. Influence of chelating ligands and As on Fe uptake/translocation*

289 Iron uptake in rice seedling was affected by chelating ligands and As significantly. Iron
290 concentration was measured both in root surfaces and plant tissues. Results show that the Fe
291 concentration was higher in rice root surfaces of control treatment (without chelating ligands)
292 while its concentration was higher in plant tissues of ligand treated nutrient solution (Fig. 5). The

293 highest Fe contents were found in tissues of rice seedlings treated with EDTA or HIDS and As.
294 Increasing Fe uptake by chelating ligands, especially EDTA and HIDS, can be explained by the
295 adsorption of As(III)-EDTA/-HIDS complex on the Fe-plaques of rice root surfaces and
296 dissociation of the complex to release of Fe(III)-EDTA/-HIDS into solution. The release of
297 Fe(III)-EDTA/-HIDS into the culture solution results in the increase of Fe uptake. Adsorption of
298 metal-EDTA to the surface of Fe oxides and dissociation of the complex and release of Fe(III)-
299 EDTA has been reported by [Nowack and Sigg \(1997\)](#).

300 Strong ligands, such as EDTA, complex with metals in natural systems. Adsorption of
301 uncomplexed EDTA on metal oxides (Fe-oxides, Al-oxides) has been studied previously
302 ([Bowers and Huang, 1985](#); [Blesa et al., 2000](#)). The EDTA has been reported to exist as complex
303 species of metals (mainly CaEDTA, ZnEDTA, and Fe(III)EDTA) in natural waters ([Xue et al.,](#)
304 [1995](#)). Dissolution reactions of Fe-oxides in the presence of metal-EDTA complexes have also
305 been reported by [Nowack and Sigg \(1997\)](#).

306

307 *3.6. Arsenic uptake/translocation affected by chelating ligands*

308 Arsenic contents in roots, shoots, and root surfaces of rice seedling were determined to
309 assess the effect of chelating ligands on As uptake. Results show that As was stored mostly in
310 roots followed by shoots and root surfaces ([Fig. 6](#)). Previous studies with rice also reported
311 higher content of As in rice roots ([Abedin et al., 2002](#)). The higher storage of As in roots and
312 lower translocation to shoots can be explained by the reduction of arsenate to arsenite in roots,
313 complexation with thiols, and sequestration in the root vacuoles ([Zhao et al., 2009](#)).

314 Formation of Fe-plaque on rice root surfaces and its effect on As uptake in rice have been
315 well explained in literature ([Liu et al., 2006](#)). Although Fe-plaque inhibits the As uptake ([Zhang](#)
316 [et al., 1998](#)), increase of the uptake of toxic and nutrient elements in plants and organisms by Fe-
317 plaque has also been reported ([Ye et al., 2001](#)). The effects of Fe-plaque on the uptake of nutrient

318 and/or toxic elements depend on the amount of Fe-plaque on root surfaces (Zhang et al., 1998).
319 Otte et al. (1989) reported higher concentration of Zn in roots of *Aster tripolium* L. coated with
320 500-2000 nmol Fe cm⁻² compared to those coated with less than 500 or more than 2000 nmol Fe
321 cm⁻². Even though the increasing amount of Fe-plaque elevates As accumulation on the root
322 surfaces, it does not affect As uptake in rice shoots. The Fe-plaque acts as “buffer” to prevent the
323 translocation of As from roots to shoots Liu et al. (2004).

324 Present study also report that the As contents in roots and shoots were higher in rice
325 seedlings grown with chelating ligands compared to those grown without chelating ligands (Fig.
326 6). Arsenic content in roots was highest when the rice seedlings were grown with HIDS while
327 the content was identical when grown with EDTA, EDDS, or IDS. The results suggest that
328 chelating ligands increased As uptake in rice root significantly, though its translocation form root
329 to shoot was not increased. The use of chelating ligands, especially the EDTA, EDDS, IDS, etc.
330 for the increase of heavy metals have been studied extensively (Jean et al., 2008; Marques et al.,
331 2008). Present study reports a better/comparable performance of HIDS to that of others studied
332 previously for the first time.

333 Arsenate has a high adsorptive affinity to Fe oxides (Zhao et al., 2009). Chelating ligands
334 solubilization/desorption As from the Fe-plaque of rice roots, and rice plant readily uptakes
335 desorbed/soluble As from the nutrient solution. The results of the present study reveal that the
336 HIDS is stronger then EDTA, EDDS, or IDS for dissolution/desorption of precipitated As. Since
337 the EDTA is not readily biodegradable, and is persistent in the environment, the biodegradable
338 HIDS would be a good alternative to EDTA in the phytoextraction/phytoremediation of As.

339

340 **4. Conclusions**

341 The use of chelating ligands in the phytoextraction of toxic metals and in the increase of
342 essential nutrient elements is not new at all. Especially, the EDTA and EDDS have been widely
343 used in agriculture for long time to serve the above purposes. The use of EDTA, however, has
344 the disadvantage that it is quite persistent in the environment due to its low biodegradability.
345 Therefore, looking for biodegradable chelating ligands is an important concern to the researchers.
346 In this study the effectiveness of HIDS for the increase of Fe bioavailability and As
347 phytoextraction was investigated, and the results were compared with those of EDTA, EDDS,
348 and IDS. The Fe limiting condition was induced by increasing the pH of the growth solution.
349 Results show that the performance of HIDS was more effective than that of other chelating
350 ligands. HIDS is a newly introduced, biodegradable and environmentally harmonious chelating
351 ligands with high chelating capability. Therefore, it would be a good alternative to the EDTA.

352

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356

357 **References**

358 Abedin, M.J., Cresser, M.S., Meharg, A.A., Feldmann, J., Cotter-Howells, J., 2002. Arsenic
359 accumulation and metabolism in rice (*Oryza sativa* L.). *Environ. Sci. Technol.* 36, 962-968.

360 Abedin, M.J., Meharg, A.A., 2002. Relative toxicity of arsenite and arsenate on germination and
361 early seedling growth of rice (*Oryza sativa* L.). *Plant Soil* 243, 57-66.

362 Alvarez-Fernandez, A., Garcia-Marco, S., Lucena, J.J., 2005. Evaluation of synthetic iron(III)-
363 chelates (EDDHA/Fe³⁺, EDDHMA/Fe³⁺ and the novel EDDHSA/Fe³⁺) to correct iron chlorosis.

364 *Eur. J. Agron.* 22, 119-130.

- 365 Anderson, M.A., Morel, F.M.M., 1982. The influence of aqueous iron chemistry on the uptake of
366 iron by the coastal diatom *Thalassiosira Weiss-ogii*. Limnol. Oceanogr. 27, 789-813.
- 367 Becker, M., Asch, F., 2005. Iron toxicity in rice - conditions and management concepts. J. Plant
368 Nutr. Soil Sci. 168, 558-573.
- 369 Berne, R.M., Levy, M.N. (Eds.), 1998. Physiology. Mosby, St. Louis, MO. USA.
- 370 Blesa, M.A., Weisz, A.D., Morando, P.J., Salfity, J.A., Magaz, G.E., Regazzoni, A.E., 2000. The
371 interaction of metal oxide surfaces with complexing agents dissolved in water. Coord. Chem.
372 Rev. 196, 31-63.
- 373 Bowers, A.R., Huang, C.P., 1985. Adsorption characteristics of polyacetic amino acids onto
374 hydrous [gamma]-Al₂O₃. J. Colloid Interf. Sci. 105, 197-215.
- 375 Bucheli-Witschel, M., Egli, T., 2001. Environmental fate and microbial degradation of
376 aminopolycarboxylic acids. FEMS Microbiol. Rev. 25, 69-106.
- 377 Chaney, R.L., Brown, J.C., Tiffin, L.O., 1972. Obligatory reduction of ferric chelates in iron
378 uptake by soybeans. Plant Physiol. 50, 208-213.
- 379 Cohen, C.K., Fox, T.C., Garvin, D.F., Kochian, L.V., 1998. The role of iron-deficiency stress
380 responses in stimulating heavy-metal transport in plants. Plant Physiol. 116, 1063-1072.
- 381 Evangelou, M.W.H., Ebel, M., Schaeffer, A., 2007. Chelate assisted phytoextraction of heavy
382 metals from soil. Effect, mechanism, toxicity, and fate of chelating agents. Chemosphere 68,
383 989-1003.
- 384 Fitz, W.J., Wenzel, W.W., 2002. Arsenic transformations in the soil-rhizosphere-plant system:
385 fundamentals and potential application to phytoremediation. J. Biotechnol. 99, 259-278.
- 386 Guerinot, M.L., Yi, Y., 1994. Iron: Nutritious, noxious, and not readily available. Plant Physiol.
387 104, 815-820.

- 388 Hansel, C.M., Fendorf S., Sutton S., Newville M., 2001. Characterization of Fe plaque and
389 associated metals on the roots of mine-waste impacted aquatic plants. *Environ. Sci. Technol.* 35,
390 3863-3868.
- 391 Helena, H., Marjatta Orama, Saarinen, H., Aksela, R., 2003. Studies on biodegradable chelating
392 ligands: complexation of iminodisuccinic acid (ISA) with Cu(II), Zn(II), Mn(II) and Fe(III) ions
393 in aqueous solution. *Green Chem.* 5, 410 - 414.
- 394 Hu, Y., Li, J.H., Zhu, Y.G., Huang, Y.Z., Hu, H.Q., Christie, P., 2005. Sequestration of As by
395 iron plaque on the roots of three rice (*Oryza sativa* L.) cultivars in a low-P soil with or without P
396 fertilizer. *Environ. Geochem. Heth.* 27, 169-176.
- 397 Hudson, R.J.M., Morel, F.M.M., 1990. Iron transport in marine phytoplankton: Kinetics of
398 cellular and medium coordination reactions. *Limnol. Oceanogr.* 35, 1002-1020.
- 399 Jean, L., Bordas, F., Gautier-Moussard, C., Vernay, P., Hitmi, A., Bollinger, J.-C., 2008. Effect
400 of citric acid and EDTA on chromium and nickel uptake and translocation by *Datura innoxia*.
401 *Environ. Pollut.* 153, 555-563.
- 402 Liu, W.J., Zhu, Y.G., Hu, Y., Williams, P.N., Gault, A.G., Meharg, A.A., Charnock, J.M., Smith,
403 F.A., 2006. Arsenic sequestration in iron plaque, its accumulation and speciation in mature rice
404 plants (*Oryza sativa* L.). *Environ. Sci. Technol.* 40, 5730-5736.
- 405 Liu, W.J., Zhu, Y.G., Smith, F.A., Smith, S.E., 2004. Do phosphorus nutrition and iron plaque
406 alter arsenate (As) uptake by rice seedlings in hydroponic culture? *New Phytol.* 162, 481-488.
- 407 Luo, C., Shen, Z., Li, X., 2005. Enhanced phytoextraction of Cu, Pb, Zn and Cd with EDTA and
408 EDDS. *Chemosphere* 59, 1-11.
- 409 Marques, A.P.G.C., Oliveira, R.S., Samardjieva, K.A., Pissarra, J., Rangel, A.O.S.S., Castro,
410 P.M.L., 2008. EDDS and EDTA-enhanced zinc accumulation by *Solanum nigrum* inoculated
411 with arbuscular mycorrhizal fungi grown in contaminated soil. *Chemosphere* 70, 1002-1014.

- 412 Morel, F.M.M., Hering, J.G., 1993. Principles and Applications of Aquatic Chemistry:
413 Complexation. John Wiley, New York, USA.
- 414 Murashige, T., Skoog, F., 1962. A revised medium for rapid growth and bio assays with tobacco
415 tissue cultures. *Physiol. Plant.* 15, 473-497.
- 416 Nowack, B., Schulin, R., Robinson, B.H., 2006. Critical assessment of chelant-enhanced metal
417 phytoextraction. *Environ. Sci. Technol.* 40, 5225-5232.
- 418 Nowack, B., Sigg, L., 1997. Dissolution of Fe(III) (hydr) oxides by metal-EDTA complexes.
419 *Geochim. Cosmochim. Ac.* 61, 951-963.
- 420 Otte, M.L., Rozema, J., Koster, I., Haarsma, M.S., Broekman, R.A., 1989. Iron plaque on roots
421 of *Aster tripolium* L.: Interaction with zinc uptake. *New Phytol.* 111, 309-317.
- 422 Pastor, J., Aparicio, A.M., Gutierrez-Maroto, A., Hernández, A.J., 2007. Effects of two chelating
423 agents (EDTA and DTPA) on the autochthonous vegetation of a soil polluted with Cu, Zn and
424 Cd. *Sci. Total Environ.* 378, 114-118.
- 425 Paul, F.B., McLaughlin, M.J., Cozens, G., Stevens, D.P., Owens, G., South, H., 2003. Plant
426 uptake of ¹⁴C-EDTA, ¹⁴C-Citrate, and ¹⁴C-Histidine from chelator-buffered and conventional
427 hydroponic solutions. *Plant Soil* 253, 311-319.
- 428 Rahman, A.M., Hasegawa, H., Mahfuzur Rahman, M., Mazid Miah, M.A., Tasmin, A., 2008.
429 Arsenic accumulation in rice (*Oryza sativa* L.): Human exposure through food chain. *Ecotoxicol.*
430 *Environ. Saf.* 69, 317-324.
- 431 Raskin, I., Smith, R.D., Salt, D.E., 1997. Phytoremediation of metals: using plants to remove
432 pollutants from the environment. *Curr. Opin. Biotechnol.* 8, 221-226.
- 433 Robinson, B., Kim, N., Marchetti, M., Moni, C., Schroeter, L., van den Dijssel, C., Milne, G.,
434 Clothier, B., 2006. Arsenic hyperaccumulation by aquatic macrophytes in the Taupo Volcanic
435 Zone, New Zealand. *Environ. Exp. Bot.* 58, 206-215.

- 436 Romheld, V., 1987. Different strategies for iron acquisition in higher plants. *Physiol. Plant.* 70,
437 231-234.
- 438 Sokubai, N., 2009. Biodegradable Chelating Agent: HIDS, Nippon Shokubai.
439 <http://www.shokubai.co.jp/eng/products/hids.html>.
- 440 Tang, D., Morel, F.M.M., 2006. Distinguishing between cellular and Fe-oxide-associated trace
441 elements in phytoplankton. *Mar. Chem.* 98, 18-30.
- 442 Tanton, T.W., Crowdy, S.H., 1972. Water pathways in higher plants: II. Water pathways in roots.
443 *J. Exp. Bot.* 23, 600-618.
- 444 Taylor, G.J., Crowder, A.A., 1983. Use of DCB technique for extraction of hydrous iron oxides
445 from roots of wetland plants. *Am. J. Bot.* 70, 1254-1257.
- 446 Tu, C., Ma, L.Q., Bondada, B., 2002. Arsenic accumulation in the hyperaccumulator Chinese
447 brake and its utilization potential for phytoremediation. *J. Environ. Qual.* 31, 1671-1675.
- 448 Vert, G., Grotz, N., Dedaldechamp, F., Gaymard, F., Guerinot, M.L., Briat, J.-F., Curie, C., 2002.
449 IRT1, an arabidopsis transporter essential for iron uptake from the soil and for plant growth.
450 *Plant Cell* 14, 1223-1233.
- 451 Wenger, K., Tandy, S., Nowack, B. (Eds.), 2005. Effects of chelating agents on trace metal
452 speciation and bioavailability. American Chemical Society, Washington DC.
- 453 Xu, X.Y., McGrath, S.P., Meharg, A.A., Zhao, F.J., 2008. Growing rice aerobically markedly
454 decreases arsenic accumulation. *Environ. Sci. Technol.* 42, 5574-5579.
- 455 Xue, H., Sigg, L., Kari, F.G., 1995. Speciation of EDTA in natural waters: Exchange kinetics of
456 Fe-EDTA in river water. *Environ. Sci. Technol.* 29, 59-68.
- 457 Ye, Z.H., Cheung, K.C., Wong, M.H., 2001. Copper uptake in *Typha latifolia* as affected by iron
458 and manganese plaque on the root surface. *Can. J. Bot.* 79, 314-320.
- 459 Zhang, X., Zhang, F., Mao, D., 1998. Effect of iron plaque outside roots on nutrient uptake by
460 rice (*Oryza sativa* L.). Zinc uptake by Fe-deficient rice. *Plant Soil* 202, 33-39.

461 Zhao, F.J., Ma, J.F., Meharg, A.A., McGrath, S.P., 2009. Arsenic uptake and metabolism in
462 plants. *New Phytol.* 181, 777-794.

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492 Figure Captions:

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494 Fig. 1: Growth of rice (*Oryza sativa* L.) affected by the pH of nutrient solution. Results are
495 presented as mean and the error bars express \pm SD ($n = 3$).

496

497 Fig. 2: Fe concentration in roots, root surfaces, and shoots of rice seedling (*Oryza sativa* L.) as
498 affected by the pH of nutrient solution. Results are presented as mean and the error bars
499 express \pm SD ($n = 3$).

500

501 Fig. 3: Fe uptake and translocation in rice seedling (*Oryza sativa* L.) as affected by chelating
502 ligand concentrations in the nutrient solution. Results are presented as mean and the error
503 bars express \pm SD ($n = 3$).

504

505 Fig. 4: Growth of rice seedling (*Oryza sativa* L.) affected by chelating ligand concentrations in
506 the nutrient solution. Results are presented as mean and the error bars express \pm SD ($n = 3$).

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508 Fig. 5: Fe concentration in root surfaces and plant tissues (roots and shoots) of rice seedling
509 (*Oryza sativa* L.) as affected by chelating ligands and arsenic in the nutrient solution.
510 As(+) and As(-) indicate with and without arsenic, respectively.

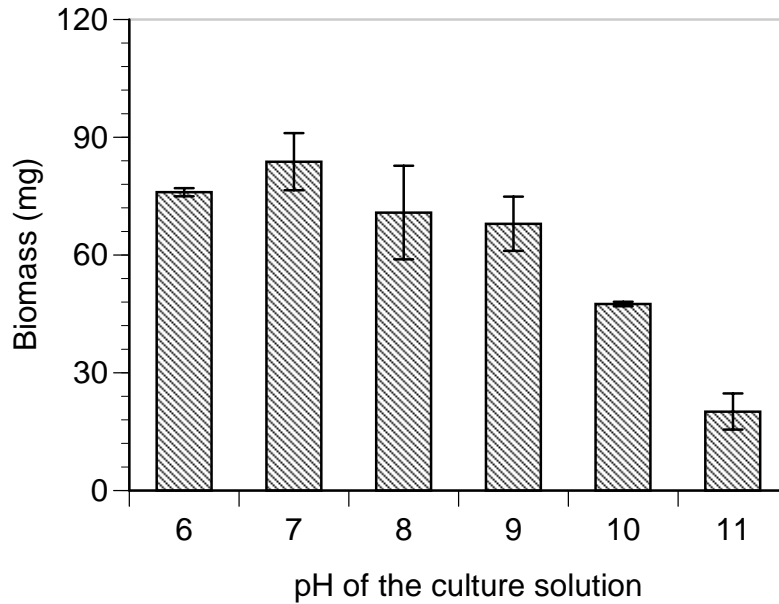
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512 Fig. 6: As concentration in roots, shoots, and root surfaces of rice seedling (*Oryza sativa* L.) as
513 affected by chelating ligands in the nutrient solution.

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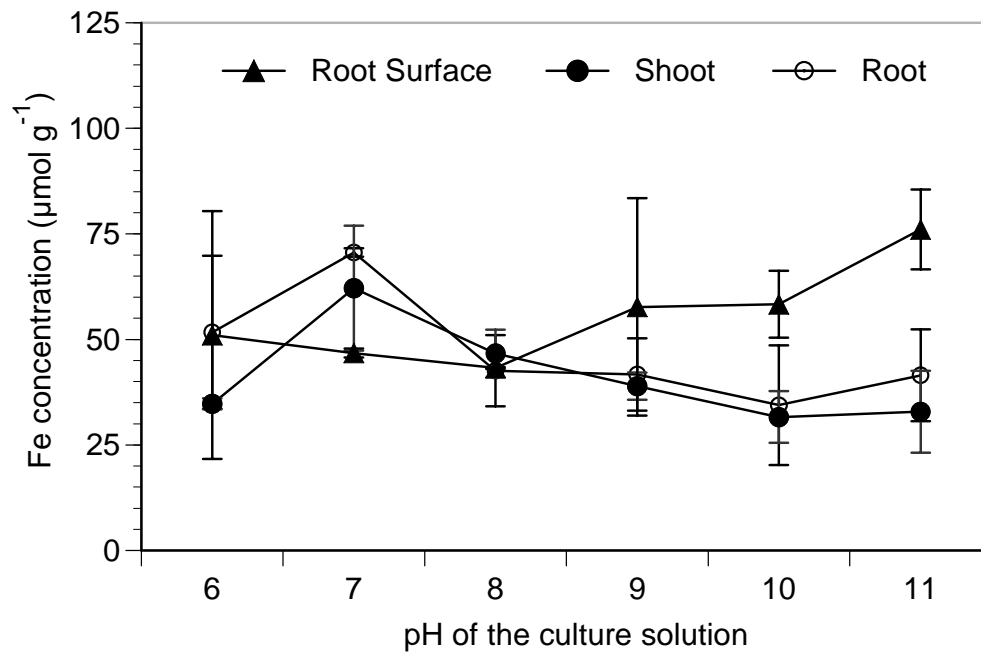
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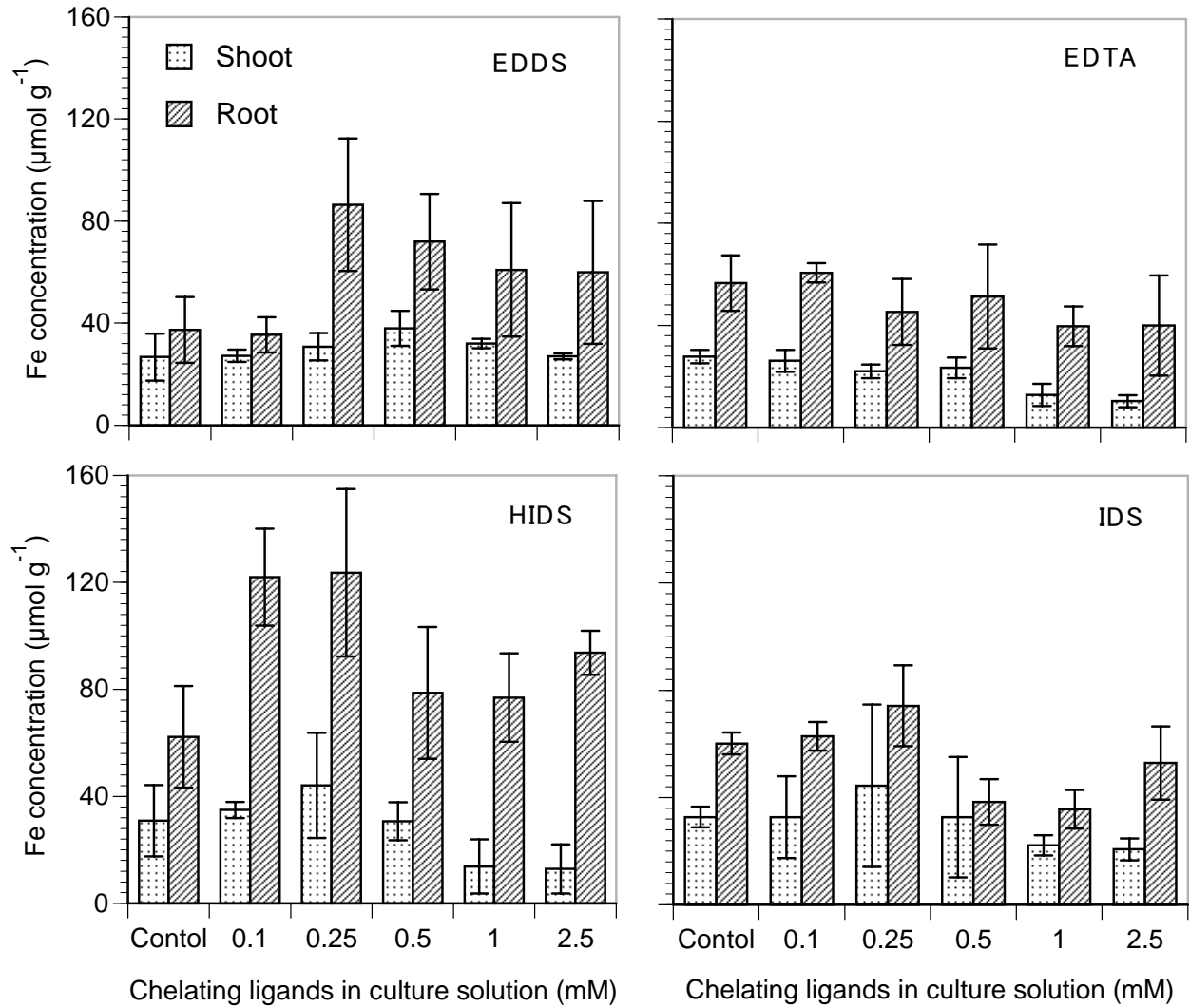
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530 Fig. 3: Fe uptake and translocation in rice seedling (*Oryza sativa* L.) as affected by chelating

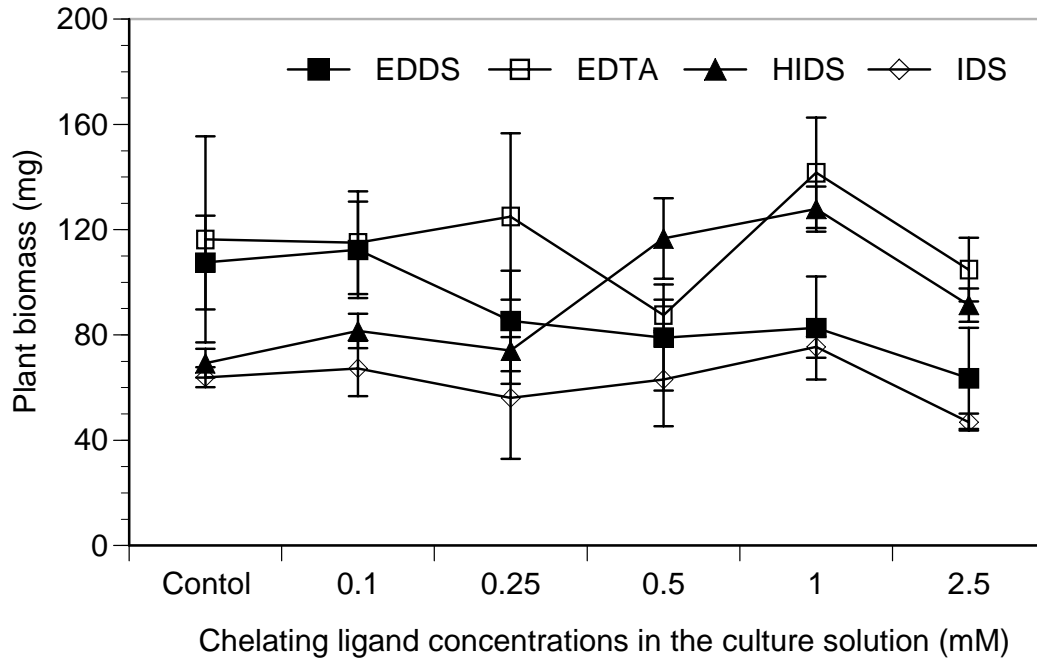
531 ligand concentrations in the nutrient solution. Results are presented as mean and the error

532 bars express \pm SD ($n = 3$).

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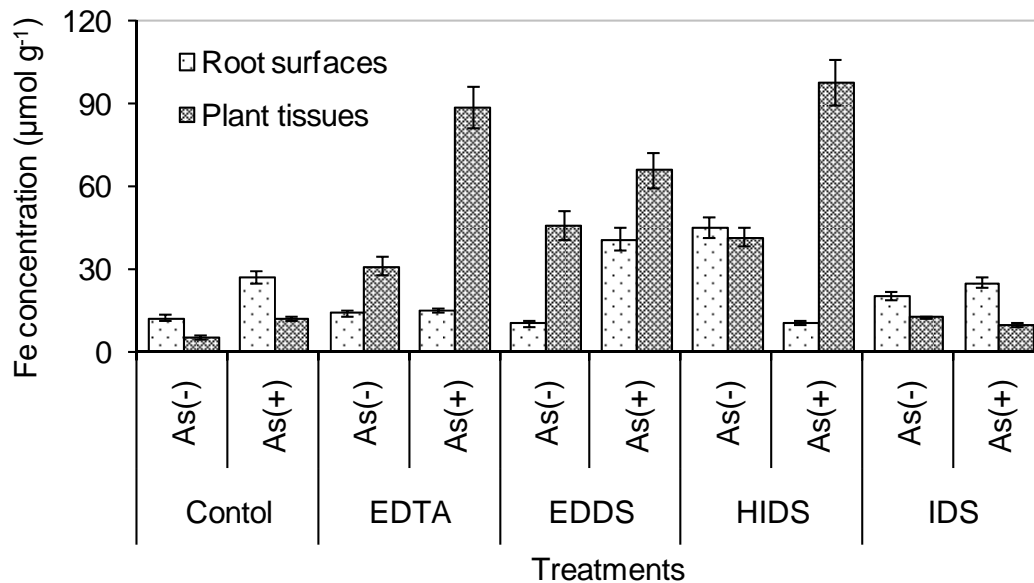


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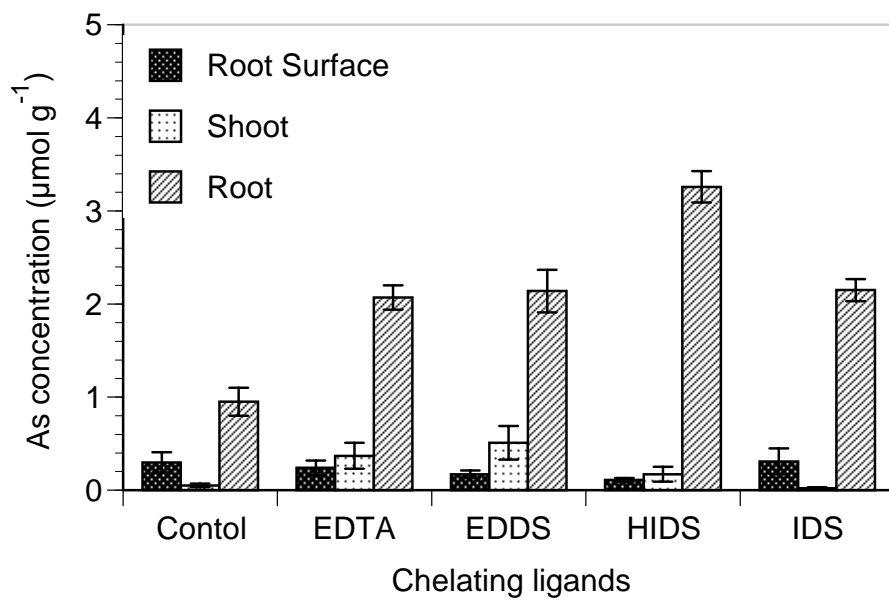
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